1	Timescale of reduction of long-term phosphorus release from sediment in lakes
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26 Abstract

28	It is important for lake management and policy to estimate the timescale of recovery from
29	long-term P release from sediment after a reduction in the external load. To provide a
30	scientific basis for this, a condensed model was elaborated, applied and evaluated in four
31	lakes. The model is based on first order kinetics, with an overall rate constant composed of
32	the rate of diagenesis of labile P (kd,2) and rate of burial of P (kb) below an active sediment
33	layer. Using the variation of P fractions in dated sediment cores, kd,2 varied from 0.0155 to
34	0.383 yr ⁻¹ , kb from 0.0184 to 0.073 yr ⁻¹ and the overall rate constant from 0.0230 to 0.446 yr ⁻
35	¹ . The active layer depths, 8 to 29 cm, and kd,2 values are within the ranges found by others.
36	The time for a 75 % reduction (t_{75}) of labile P in the active layer is 60 years in Lough Melvin,
37	3 in Ramor, 33 in Sheelin and 41 in Neagh, although P release is only important in Ramor
38	and Neagh. Combining the kd,2 values with other estimates (mean 0.0981 yr ⁻¹ , median
39	0.0426; n=14) produces a t_{75} value of less than 14 and 33 years. A review of other models
40	indicates a timescale of one to two decades and from lake monitoring also of one to two
41	decades. It is desirable to estimate the timescale directly in all lakes if sediment P release is
42	important, but, generally, it should take between one and three decades.
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45	Key words
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47	Recovery
48	Model
49	Diagenesis
50	Burial

52 **1** Introduction

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1.1 Recovery of lakes from excessive phosphorus concentrations

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Around the world, the recovery of lakes from excessive phosphorus concentrations has been
a research and a practical topic in limnology since at least the 1960s, with lake phosphorus
models a key element (Prairie 1989; Brett & Benjamin 2008). The models have generally
been used to set external P loading targets that support desirable lake water P concentrations
and biological conditions, as outlined quite early in the field (Dillon & Rigler 1975).

61

62 This structure is still relevant and used in the context of the European Union's Water Framework Directive (2000/60/EC), in that lakes and other water body types are required to 63 achieve biological standards, usually set at Good Status. The status, one of five classes from 64 65 High to Bad, is determined by the deviation of the phytoplankton, other aquatic flora, benthic invertebrate fauna and fish fauna from undisturbed conditions (Quevauviller et al. 2008). I 66 67 some lakes, it can take time to reach a lake water P concentration that supports Good Status and for the biological properties on which status is based to respond. Therefore, the 68 69 timescale of achieving Good Status should be considered during development of a River 70 Basin Management Plan. Specifically, the deadline for achieving Good Status can be extended due to, amongst other things, the natural delay in the recovery processes such as the 71 timescale of reduction in internal P load in some lakes. 72

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74 1.2 Internal phosphorus load

76 The failure of models to predict the water column TP concentration in some lakes was observed quite early on and the process of enhanced P release from sediment, the internal P 77 load, identified (Bengtsson 1975; Ahlgren 1977; Larsen et al. 1979); see also the compilation 78 79 by Jeppesen et al. (1997). For example, Cullen & Forsberg (1988) found that 28 out of 45 lakes that had reductions in their external P load had no or very little change in trophic state. 80 The internal P load was one of the factors that caused delayed recovery, along with the 81 degree of reduction in the external load, growth limiting factors other than P (mainly 82 nitrogen) and hydraulic residence time. 83

84

When applied to a change in external load, the models underestimate the TP concentration by 85 not including varying internal P load or having a single value for the P 86 87 retention/sedimentation coefficient (Nürnberg 1984; Nürnberg 1998). A change in the value of the P retention/sedimentation coefficient has been documented. For example, Müller et al. 88 (2014) observed that the sedimentation coefficient in four deep lakes was constant up to a 89 critical lake TP concentration, decreasing above it; typically from 0.45 to 0.1 yr⁻¹ in three of 90 the lakes, 0.7 to 0.4 in the other. Rippey & Anderson (1996) used the sedimentary record to 91 reconstruct the P loading and dynamics in a small lake and found that the retention and 92 sedimentation coefficients decreased rapidly as the lake became eutrophic; retention 93 coefficient decreased from 0.67 to 0.59 yr⁻¹ and sedimentation coefficient from 0.54 to 0.08 94 vr^{-1} . 95

96

97 In those lakes with a considerable internal P load, whether the retention/sedimentation
98 coefficient changes over time or not, the timescale of reaching a new steady-state
99 concentration in static models is also influenced by loss of P through the outflow and so by
100 the hydraulic residence time (τ_w). Typically, the time for the concentration to reach 95 % of

101 the way to the new steady-state value (t₉₅) is used and it is $3.00\tau_w$; for t₉₀ it is 2.30, t₇₅ 1.39 τ_w 102 and t₅₀ 0.693 τ_w (Chapra 1997, pp. 60-61). The timescale could be longer, depending on 103 how quickly the internal P load decreases and how the retention/sedimentation coefficient 104 responds to reduction in external load and lake water concentration. There is also added 105 uncertainty to the timescale, as some biological groups on which status is based, particularly 106 phytoplankton and macrophytes, take longer to change than the lake water P concentration 107 (Jeppesen et al. 2005; Verdonschot et al. 2013; McCrackin et al. 2016).

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109 1.3 Sediment phosphorus models

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111 Changes of internal P load have been incorporated into only a few lake models. Examples of 112 condensed lake-sediment models are Chapra & Canale (1991), who linked the amount of P 113 recycling to the oxygen regime, and Jensen et al. (2006). The latter needs a considerable 114 amount of input data and uses four model coefficients, but it did reproduce the seasonal water 115 column P dynamics in 16 shallow eutrophic lakes.

116

Focussing on describing the long-term release of P from sediment, as this is key to how long 117 it takes a lake with considerable internal load to reach a new steady state after a decrease in 118 external load, the few models available are full diagenetic ones. That of Katsev et al. (2006) 119 indicates that the long-term P efflux depends on immobilizing P, in conjunction with 120 permanent burial, which is different from the factors that influence the well-known seasonal 121 changes. Hupfer & Lewandowski (2008) also concluded that long-term P retention in 122 sediment is controlled by P fixation in reduced layers below the oxic surface layer, rather 123 than by the relatively small amount of transient and time-variable P fixation at the oxic 124 sediment surface. 125

127	Lewis et al. (2007) also constructed a diagenetic model of P in sediment and applied it
128	successfully to describing the observed release rate of P from sediment and its seasonal
129	variation in Lake Onondaga. They also presented a simple mass-balance model for P in
130	sediment that uses only two factors, burial of P in the sediment and diagenesis of labile P. As
131	this model requires relatively little work to estimate the timescale of recovery from long-term
132	release of P from the sediment in a lake, the aim of this investigation was to evaluate it during
133	application to four lakes. As the model predicts future changes, they cannot be compared to
134	observations and so the evaluation is based on assessing the behaviour of the model,
135	comparing the values of the model coefficients and the timescales to the measurements and
136	observations of others.
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138	2 Material and methods
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140	2.1 The model
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142	The model used was suggested by Lewis et al. (2007) and is a box (Continuously Stirred
143	Tank Reactor) model of the rate of change of labile P in an active layer of sediment, labile P
144	taken to be responsible for the long-term release of P from sediment. It describes the
145	diagenesis and burial of P and, as the diagenesis rate constant is estimated using the concepts
146	and methods of Penn et al. (1995), it could be called the Lewis/Penn model.
147	
148	It is depicted in Fig. 1. Freshly deposited P consists of labile and refractory P, and labile P
149	undergoes slow diagenesis by first order kinetics in the active sediment layer; the Lewis et al.
150	(2007) model includes fast diagenesis, but it is ignored here as the focus is the longer

151 timescale. Some mineralized phosphate is released from the sediment, some immobilized in the active layer and some remains in the active layer. Refractory P, phosphate and 152 immobilized phosphate leave the active layer through permanent burial. 153

154

The rate of change of labile P in the active sediment layer is described using first order 155 kinetics: C = exp(-kt), where C is the concentration of labile P and k the rate constant of 156 157 change of C with time t. k describes or is composed of two processes, diagenesis and burial. Using the addition of rate constants/residence times (Lerman 1979, pp. 5), $k = kd_{2} + kb_{3}$, 158 159 where kd,2 is the slow diagenesis rate constant and kb the rate constant for burial; Lewis et al. (2007) describe the rate constant for burial as ω /Zactive (ω the mean sediment 160 accumulation rate in the active sediment layer and Zactive the active layer depth), so kb= 161 ω /Zactive. 162 163

164 The three model coefficients (ω , Zactive and kd,2) are estimated using the concentrations of operationally-defined sediment P fractions in a dated sediment core. The P fractions of 165 Hieltjes & Lijklema (1980) are NH4Cl-P, NaOH-P, HCl-P and residual-P, modified to 166 differentiate NaOH-P into reactive NaOH-P and non-reactive NaOH-P (Furumai & Ohgaki 167 1982); see also Ruban et al. (1999). 168

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Penn et al. (1995) estimated labile P by defining two sediment P fractions. Fraction A 170

consists of the sum of NH4Cl-P, reactive NaOH-P and non-reactive NaOH-P and comprises 171

labile and refractory P; Fraction B is HCl-P and residual-P and is taken to be all refractory. 172

As all P below the active sediment layer is assumed in the model to be refractory, the ratio of 173

174 Fraction A to B there (α) is constant and the ratio is used to estimate the refractory

component of Fraction A; refractory Fraction A = α Fraction B. Labile P is then Fraction A 175

176	minus the refractory component of Fraction A. Zactive is the depth in the sediment where α
177	becomes constant and ω the mean sediment accumulation rate in the active layer.

To calculate kd,2, Penn et al. (1995) estimated the labile P concentration in freshly deposited sediment by assuming that the ratio of labile to refractory P in fresh sediment (r) is constant, so labile P in fresh sediment is r x refractory P concentration. They found 49 % of the total P in sediment trap material in Lake Onondaga was labile P, so r = 49/51 = 0.96. The integrated first order kinetic equation, ln(labile P/labile P in fresh sediment) = kd,2t, is used to estimate kd,2. Using different values of r does not change the value of kd,2, although it does the concentration of labile P in fresh sediment.

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187 2.2 Site description

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The model was applied to four lakes with one sediment core from each of Lough Melvin, Ramor and Sheelin and five from Lough Neagh (Table 1). We avoided coring the deepest point in Melvin, Ramor and Sheelin, choosing a slightly shallower site that should better represent the sedimentary basin as it is less affected by sediment focusing. Five cores were retrieved from across the Lough Neagh sedimentary basin to help establish the variability of the model coefficients and timescale. It has a relatively flat basin, with most of the lake depth between 9 and 13 m (Carter 1993).

196

197 Based on the annual mean lake water TP concentration, calculated using results from a

198 variety of papers, reports and the Environmental Protection Agency of Ireland's EDEN

199 system, and applying the widely used trophic state classification (Nurnberg 1996), the trophic

200 history of the lakes is as follows.

202	Lough Melvin is a Special Area of Conservation under the European Union's Habitats
203	Directive (92/43/EEC) because of its oligo- to meso-eutrophic conditions, with Littorella,
204	Lobelia, Isoetes and Atlantic salmon present, and its terrestrial vegetation. The lake has been
205	mesotrophic (TP<30 μ g L ⁻¹) over the 2007-2015 period at least. Ramor has been eutrophic
206	(>30 μ g L ⁻¹) since at least 2000 but with the P concentration decreasing since 2006. Sheelin
207	changed from meso- to eu-trophic in the mid-1970s and since then has varied between the
208	two trophic states, improving and maintaining mesotrophic status since 2008.
209	
210	A considerable amount is known about Lough Neagh, regarding its limnology and changes in
211	key variables (Wood & Smith 1993; Stronge et al. 1998; Gibson et al. 2000). It is a Special
212	Protection Area under the Habitats Directive and a Ramsar site because of its aquatic
213	vegetation, fringing wetlands, invertebrates, waterfowl and the endangered pollan
214	(Coregonus autumnalis), and it also supports a large commercial eel fishery. A summary of
215	the trophic history as reconstructed by Foy et al. (2003) and Bunting et al. (2007) is as
216	follows. Productivity began to increase after 1880, more rapidly after 1950 as a result of
217	increased external P loading. Sometime during the 1960s, the internal P load increased
218	greatly (Gibson et al. 2001) and since then increases in external nitrogen loading have largely
219	been responsible for further increases in productivity. Currently, the lake is hypereutrophic,
220	based on annual mean TP (>100 μ g L ⁻¹) and chlorophyll <i>a</i> (>25.9 μ g L ⁻¹) concentration and
221	eutrophic based on TN (>650 µg L ⁻¹) (Nurnberg 1996).
222	

223 2.3 Methods

One metre sediment cores were retrieved (Mackereth 1969), stored at 4 °C and sectioned
using the covered slice method to determine wet density (Hilton et al. 1986) into 1 cm slices.
The samples were purged with nitrogen and capped (Lukkari et al. 2007). The analysis of
sediment P fractions started the following day, with 1.5 g subsamples of wet sediment
transferred into 50 mL polypropylene tubes. Dry weight (105 °C), loss on ignition (550 °C)
and carbonate content (1000 °C) were measured (Dean 1974) on another subsample.

231

Persulphate digestion (Eisenreich et al. 1975) was used to determine TP in the NaOH extract 232 and phosphate was measured in all extracts by solution spectrometry (Murphy & Riley 1962). 233 234 Residual-P was determined by sequential addition of HF, HNO₃ and HClO₄ in Teflon beakers (Bock 1979), with the ignition method (Andersen 1976; Ostrofsky 2012) used for the Lough 235 Neagh cores. The NaOH and HCl extracts were neutralized and a twentyfold dilution 236 generally needed before phosphate analysis. The precision of results (standard deviation, mg 237 $P g^{-1} DS$) was established using replicates (n=10-14) of a sediment sample from Lough 238 Neagh; NH4-Cl-P, 0.000238; reactive NaOH-P 0.0493, non-reactive NaOH-P, 0.0264; HCl-239 P, 0.0376; residual-P, 0.0214. The mean ratio (\pm SE) of the sum of the five P fractions to the 240 total P concentration determined independently by HF/HNO₃/HClO₄ digestion in 13 241 replicates was 1.00 (±0.0273). 242

243

The sediment cores were dated by measuring ²¹⁰Pb, ²²⁶Ra, ¹³⁷Cs and ²⁴¹Am by direct gamma
assay in the Environmental Radiometric Facility at University College London, using a
ORTEC HPGe GWL series well-type coaxial low background intrinsic germanium detector
(Appleby et al. 1986; Appleby et al. 1992). Lead-210 dates were calculated using the CRS
(constant rate of ²¹⁰Pb supply) dating model and the ¹³⁷Cs activity-depth profile used to
support them (Appleby & Oldfield 1978; Appleby 2001).

- 250
 251 **3** Results
 252
- 253 3.1 Basic results
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The concentrations of NH4Cl-P, reactive NaOH-P, non-reactive NaOH-P, HCl-P and 255 residual-P are given in Table 1 to 8 in the Supplementary material and a summary is as 256 follows. The NH4Cl-P concentrations were zero or very small in most samples, the highest 257 being 0.002 to 0.056 mg P g⁻¹ DS (dry sediment) in the 0-1 cm layers. Reactive NaOH-P was 258 the most variable with a range of 0.1 to 2.0 mg P g^{-1} DS over all the lakes and the 259 concentration generally decreased with depth in the sediment. Non-reactive NaOH-P had the 260 next largest range, 0.1 to 1.5 mg P g⁻¹ DS, while the ranges for HCl-P and residual-P were 261 similar at 0.2 to 0.7 and 0.2 to 0.8 mg P g^{-1} DS, respectively. 262 263 These values are within the ranges found by Ostrofsky (1987) in the surficial sediment of 62 264 lakes in eastern North America; mean and range of reactive NaOH-P are 0.82 and 0.14-5.88 265 mg P g⁻¹ DS with HCl-P 0.27 and 0.04-0.74. The concentrations of all fractions except 266 residual-P in the calcareous sediment of Lake Onondaga are different to the ranges found in 267 the Irish lakes; NH4Cl-P and HCl-P are higher, with a range up to 0.4 and 1.6 mg P g-1 DS, 268 respectively, and reactive and non-reactive NaOH-P lower, up to 0.25 for both fractions 269 (Penn et al. 1995). In two sediment cores from another calcareous lake, Lake Balaton, 270 Herodek & Istvanovics (1986) found low concentrations of all fractions; NH4Cl-P was <0.05 271 mg P g⁻¹ DS, reactive NaOH-P 0.05-0.1, HCl-P 0.3-0.4 and residual-P 0.1 to 0.2; the TP 272 concentration was also low at 0.6 mg P g⁻¹ DS. Many more results are available for the 273 sediment fractionation scheme of Psenner et al. (1988). 274

The dating of the cores using the CRS model, supplemented by ¹³⁷Cs, provided good 276 chronologies (Table 9 to 16 in the Supplementary material) and there was only a minor 277 difficulty with one. In Lough Sheelin (SHE5), the unsupported ²¹⁰Pb activity is relatively 278 constant in the 0-16 cm sediment layer, as are the sedimentary P fraction concentrations (see 279 Fig. 2). This mixed layer may have been the result of redistribution of the sediment by wind 280 in Spring 2018. The indirect evidence for this is the large areas of coloured/turbid water 281 observed in the north-eastern part of the lake during a six week period of strong easterly 282 283 winds in March-April and, given the area and maximum depth (Table 1), Lough Sheelin is prone to the random redistribution of sediment by wind (Hilton 1985). 284

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286 3.2 Sediment phosphorus fractions

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The key P fraction used in the Lewis/Penn model is the variation with depth of Fraction A as 288 a percentage of the TP concentration and this is presented in Fig. 2 and 3. Fraction A and the 289 labile P concentration do decrease steadily with depth in the sediment of all the cores and 290 labile P does become zero (Fig. 1 & 2). The clearest steady reduction of labile P with depth 291 is in Melvin and Lough Neagh cores LN11, 17, 18 and 19. This general trend is irregular in 292 Ramor, is interrupted by lower values below the trend in Lough Neagh core LN15 and there 293 294 is a mixed layer in Sheelin. The irregular trend in Ramor and decline below the trend between 9 and 17 cm in LN15 is caused largely by changes in reactive NaOH-P. In Ramor, 295 as TP is much less variable (Fig. 1 in the Supplementary material), the changes in Fraction A 296 297 and labile P do not appear to be the result of unusual sedimentation at the coring site or to errors in the chemical analysis. This is not the case in LN15, as TP is also below the trend. 298

As the dry weight is higher and LOI lower in this 9-17 cm interval, there may have beensome change in the nature of depositing sediment over this time period at the site.

301

302 The mixed 0-16 cm sediment layer in Sheelin means there is some uncertainty in the sediment accumulation rate and active layer depth. While the evidence suggests this layer 303 was the result of redistribution of sediment during a prolonged windy period, the layer is 304 quite deep and it is not known if it occurs over the sedimentary basin. If it does, then the core 305 would yield typical results for the lake. If it doesn't, then results from below 16 cm depth 306 307 may better characterize the long-term behaviour of labile P. We used all the results from the sediment surface to the depth at which Fraction A becomes constant to calculate ω and 308 309 Zactive; the diagenesis rate constant can only be derived using results below the mixed layer. Whether the mixed layer should be included or omitted only affects kb and the values are 310 0.0625 (see Table 2) and 0.0958 yr⁻¹, respectively. The diagenesis rate constant (0.383 yr⁻¹; 311 see Table 2) is four times higher than the higher kb value, so the overall rate constant is only 312 slightly affected by whether the 0-16 cm layer is included or not. 313 314

315 3.3 Model coefficients and rate constants

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Although these features detract from the steady decrease of labile P with depth in three of the
cores, they still allowed the three model coefficients and two rate constants to be estimated
(Table 2; Fig. 4 & 5).

320

321 Zactive varies between 15 and 28 cm in Melvin, Ramor and Sheelin and between 8 and 29
322 cm in the five Lough Neagh cores. Three of the Neagh cores have values between 21 and 29

323 cm, a range similar to the other lakes, while the two cores with lower values (LN11, 8 cm and 324 LN19, 10 cm) also have lower ω values (0.229, 0.145 cm yr⁻¹).

325

kb varies from 0.00753 yr⁻¹ in Melvin to 0.0625 in Sheelin and from 0.0145 to 0.0286 yr⁻¹, with a mean (\pm SE) and CV (standard deviation/mean, expressed as a percentage) of 0.0184 yr⁻¹ (\pm 0.00263) and 32.0 %, respectively, in the five Lough Neagh cores. If the value for LN11 (0.0286) is omitted, as it is almost twice the other values, the reproducibility in the Lough Neagh cores improves, decreasing the CV to 10.4 %.

331

kd,2 values were always able to be derived from the kinetic plot, even with the irregularities
in labile P noted. The data points that were omitted when calculating kd,2 are identified in
Fig. 4 and 5. The 0-16 cm mixed layer in Sheelin is quite deep and unusual, but the 13 data
points below the layer form a clear first order decay and so we have no reason to discard the
kd,2 value.

337

The kd,2 values vary from 0.0155 yr⁻¹ in Melvin to 0.383 in Sheelin and from 0.138 to 0.184 yr⁻¹, with a mean (\pm SE) and CV of 0.0165 yr⁻¹ (\pm 0.000867) and 11.7 %, respectively, in the five Lough Neagh cores. The replication of kd,2 in the Lough Neagh cores is good, given that ω and Zactive are more variable (CV 45.5 and 49.1 %, respectively; Table 2).

342

The overall rate constant (k) is the sum of kd,2 and kb (Table 2). k varies from 0.0230 yr⁻¹ in Melvin to 0.446 in Sheelin and from 0.0302 to 0.0470 yr⁻¹, with a mean (\pm SE) and CV of 0.0349 yr⁻¹ (\pm 0.00310) and 19.9 %, respectively, in the five Lough Neagh cores.

The relative contribution of burial to the total loss of labile P from the active sediment layer is kb/k, expressed as a percentage, and this is shown in Table 2. Both diagenesis and burial are important; the contribution of burial varies from 14 % in Sheelin to 33 in Melvin and from 44 to 61 % with a mean (\pm SE) and CV of 51.9 % (\pm 2.91) and 12.6 %, respectively, in the five Lough cores.

352

353 3.4 Timescales

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The time for the labile P concentration in the active sediment layer to reach 50 (t_{50}), 75 (t_{75}) 355 and 90 (t₉₀) % of the way to a new steady-state through diagenesis and burial is determined 356 by the overall rate constant and these are shown in Table 2. Taking t₇₅ as an indicator of the 357 time it takes for long-term sediment P release to become unimportant to the lake P budget, 358 then it is 3 years in Sheelin, 33 in Ramor and 60 in Melvin, with a mean (±SE) and CV of 359 40.8 years (±2.85) and 15.6 % in Lough Neagh. Consideration of the relative importance of 360 internal and external loads would help to indicate whether t_{75} or t_{90} best represents the 361 timescale of recovery. However, a lake model that describes the rate of change of lake 362 concentration with reductions in external and internal loads is desirable in order to be more 363 364 precise.

365

366 4 Discussion

367

The aim of the investigation was to apply and evaluate the Lewis/Penn model to four lakes in order to estimate the timescale of recovery from long-term sediment P release. As the model predictions cannot be compared to observations in the lakes, the model is evaluated by assessing its behaviour, comparing the values of the model coefficients and timescales to themeasurements and observations of others.

373

374 4.1 Behaviour of the model

375

The model was applied to the four lakes without difficulty. Judgement only had to be applied to occasional reactive NaOH-P concentrations that departed from the general trend in part of a core and to the mixed sediment layer in Lough Sheelin, but values for the three model coefficients were always able to be estimated.

380

Based on the five cores from Lough Neagh, the good precision of the burial rate constant (CV
32.0 %), diagenesis rate constant (11.7) and overall rate constant (19.9) estimates indicates
that the model is well poised, even though the sediment accumulation rate and active
sediment layer depth are more variable, with CVs of 45.5 and 49.1 %.

385

The model indicates that both diagenesis and burial of P are important in the lakes, with burial as a proportion of the total loss of labile P from the active layer varying from 14.0 % in Sheelin to 51.9 % in Lough Neagh (Table 2). Katsev et al. (2006) found that immobilization of phosphate and permanent burial of all forms of P were the main influences on long-term P release from sediment in their full diagenetic model, at least when redeposition of released P was omitted.

392

Carey & Rydin (2011) analyzed the P-depth patterns in sediment cores from 94 lakes and
found that the concentration stabilized at a depth of 16 cm in half of the eutrophic lakes.
Their view was that the potentially mobile P in sediment was between the surface "to the

depth at which P diagenesis processes have stabilized and where the remaining P becomes
permanently buried...". This concept is similar to that used in the Lewis/Penn model and it
has also been employed by Dittrich et al. (2009) and Horppila et al. (2017) to estimate
potential P release from sediment.

400

The three model coefficients were all measured in dated sediment cores and only the ratio of labile to refractory P in fresh sediment was assumed. Usefully, varying its value does not alter the diagenesis rate constant but the ratio is assumed to be constant. If it did vary over time, for example with the composition of depositing phytoplankton, it would affect the estimation of labile P in fresh sediment from the refractory P concentration (Fig. 1). However, the refractory P concentration varies approximately by a factor of two in each core, so small changes in the ratio should have little effect on the labile P concentrations.

If the timescale of long-term reduction of labile P only is needed, the amount of work in 409 applying the model can be reduced. Only the variation of Fraction A and Fraction B with 410 depth in the sediment core is needed to estimate Zactive, ratio of Fraction A to Fraction B, 411 labile P, refractory P and labile P in fresh sediment. So, as Fraction A consists of the sum of 412 NH4Cl-P, reactive NaOH-P and non-reactive NaOH-P, it could be measured in one step by 413 extracting the sediment with 0.1 M NaOH and determining the TP concentration. Similarly, 414 415 Fraction B consists of the sum of HCl-P and residual-P and it could be measured by determining residual-P on the residue from the NaOH extraction. All the other calculations 416 would be as described in Material and methods. 417

418

419 4.2 Model coefficients

The model is further evaluated by comparing the values of two model coefficients to
measurements made by others; the sediment accumulation rate is not included as it is lake
specific.

424

The active sediment layer depth in the four lakes varied between 15 and 28 cm in Melvin,
Ramor and Sheelin and between 8 and 29 cm in the five Lough Neagh cores (Table 2).
Similar values for the depth of sediment over which most of the diagenesis of labile P occurs
have been found using a variety of other methods (Table 3). They generate a range of 10 to
30 cm, with a typical value of 15, which is also similar to the stabilization depth for P of 16
cm in sediment cores found in many eutrophic lakes by Carey & Rydin (2011).

431

The diagenesis rate constant of labile P in the four lakes varied from 0.0155 yr⁻¹ in Melvin to 432 0.383 in Sheelin, with a mean of 0.0165 in Lough Neagh (Table 2). Penn et al. (1995) 433 measured the value in Lake Onondaga and collated other values, and these and others derived 434 using a range of methods have been reviewed and are presented in Table 4. The range is 435 quite large $(0.00701 \text{ to } 0.383 \text{ yr}^{-1})$, but the method does not seem to influence the magnitude; 436 the two highest values were obtained by slightly different methods, 0.383 yr⁻¹ in Sheelin by 437 the sedimentary P fractionation and 0.404 by the change of phosphate mono- and di-esters 438 and polyphosphates with depth in a sediment core. 439

440

The smallest values tend to be found in lakes with low lake water TP concentrations or that are classified as less than eutrophic (Table 4). This difference was tested using a single factor ANOVA with two groups, one (n=7) with an annual mean TP concentration <30 ug L⁻¹ or classified as less than eutrophic and the other (n=7) >30 ug L⁻¹ or eutrophic or greater. Logarithmic transformation achieved normality (Shapiro-Wilks p>0.72) and homogeneity of variance (Levene's p=0.468). The ANOVA is not conventionally significant (p=0.0788; F=3.52), although close, and it has a Type II error of $0.20 < \beta < 0.50$ for a 0.05 (Zar 2010, pp.114-118). If a Type I error a of 0.10 is used (Zar 2010, pp. 78-80), the ANOVA is significant with a Type II error of $0.10 < \beta < 0.20$. So, based on a Type II error of < 0.20, as is commonly the case (Zar 2010, pp. 79), then there is a difference in the diagenesis rate constant between lakes with low (mean kd,2 0.025 yr⁻¹) and high (0.078 yr⁻¹) TP concentrations.

453

While many may consider this sufficient evidence that the diagenesis rate constant is higher 454 in eutrophic lakes, in order to estimate a typical value at this stage we calculate mean and 455 median values. However, so that the values relate to a decadal timescale, we omit the four 456 457 values derived by calibration of models for seasonal changes (Table 4), which gives an average (n=14) of 0.0981 yr⁻¹ and median of 0.0426 yr⁻¹. Two of the remaining values are 458 much higher than the rest (0.383 in Sheelin, 0.404 in Sonderby) and if they are omitted, the 459 mean is 0.0489 and median 0.0341 yr^{-1} . It is always desirable to measure the diagenesis rate 460 constant in a lake where internal P load is important so that a reliable estimate is available of 461 how quickly this load reduces after external load reduction, but the mean and median could 462 be taken to be typical values. 463

464

465 4.3 Timescales

466

The overall rate constant describes how quickly the concentration of labile P in the active
sediment layer decreases to a new steady-state value due to the combination of diagenesis
and burial. If a reduction of 75 % is used, that timescale is 3 years in Sheelin, 33 in Ramor
and 60 in Melvin, with 41 in Lough Neagh (Table 2). If the average (0.0981 yr⁻¹) and median

471 (0.0426) values from Table 4 are used to represent typical behaviour, the contribution from
472 burial would not be included and so the time period would be overestimated. These values,
473 nonetheless, give timescales (t₇₅) of less than 14 and 33 years.

474

Other models have been used to estimate the timescale of reduction of internal P load. The 475 diagenetic model of Katsev et al. (2006) indicates 10 to 20 years and that of Lewis et al. 476 477 (2007) produced t₉₀ and t₉₅ values of 16 and 25 years. The lake sediment box model of Schippers et al. (2006) produces a t₈₀ value of 70 years, based on coefficients appropriate for 478 479 a shallow, low hydraulic residence time lake. The diagenesis rate constants derived by Wilson et al. (2010) for three acidic ponds were included in Table 4 and they reported t₅₀ 480 values of 24 to 99 years, depending on the pond and model used, with an outlier of 546; as 481 these are oligotrophic lakes (annual mean TP all less than 8 μ g L⁻¹) and so have little or no 482 internal load, the values are not relevant for eutrophic lakes. The model of Chapra & Canale 483 (1991) requires records for the oxygen concentration in the hypolimnion, but applying it to a 484 simulation of a major reduction in external load to Lake Shagawa, t₅₀, t₇₅ and t₉₀ values of 24, 485 486 48 and 80 years were derived. Horppila et al. (2017) estimated that substantial internal load 487 would continue for approximately 25 years in shallow, hypereutrophic Lake Tuusulanjarvi, based on the difference in P concentration between the surface and the stabilization depth in a 488 489 dated sediment core.

490

While the timescales produced by these models used different criteria for the degree ofchange, overall the values are one to two decades, occasionally larger.

493

The results from long-term monitoring of lakes, where sediment P release was important andreduction in external load had occurred, provide direct evidence of the timescale of approach

496 to the new steady-state concentration. Jeppesen et al. (2005), based on 35 lakes throughout Europe, found that sediment P release delayed the recovery, especially in eutrophic (shallow) 497 lakes. The evidence was that it took between 10 and 15 years to reach the new steady-state 498 499 concentration. The external P load in Loch Leven was reduced by approximately 50 % between 1985 and 1995, but May et al. (2012) and Spears et al. (2012) found that the lake 500 concentrations were largely determined by the internal load; the P retention coefficient was 501 0.44 in 1975. 0.61 in 1985, 0.12 in 1995 and 0.15 in 2005, showing that the increase in 502 internal load began after reduction in the external load and that release from the sediment was 503 504 still important more than ten years after external load reduction. Shatwell & Kohler (2019) described the change in low hydraulic residence time (0.11-0.23 yr) Lake Mugglesee after the 505 external N and P loads were reduced by 79 and 69 %. While the N concentration in the lake 506 507 reduced proportionally to the external load reduction, this was not the case with P and the 508 internal P load was still important more than 20 years after external load reduction. 509

510 The timescales observed during these monitoring programmes varies from three to 20 years,511 but the majority are between one and two decades.

512

In summary, the weight of evidence from the four study lakes (3-60 years), the general
diagenesis rate constant (< 14-33), other models (one to two decades) and from lake
monitoring (one to two decades) is a timescale of reduction of internal P load from one to
three decades. While this is strong evidence for the general timescale, it is always desirable
to estimate it directly in lakes that have a sizeable internal P load.

518

519 4.4 Application of the model to the four lakes

521 The model was applied to Lough Melvin, Ramor, Sheelin and Neagh and, while timescales 522 for the reduction of labile P in the active sediment layer were derived, they are only relevant 523 to lake management if long-term sediment P release is an important element of the P budget 524 in the lake.

525

526 The timescale of recovery after reduction of the external P load is determined by the 527 hydraulic residence time and the rate of decrease of internal P load, but only if the internal 528 load is an important component of the lake P budget. Lakes with a short residence time will 529 respond more quickly, as loss of nutrients via the outflow will be important.

530

The best evidence that sediment P release is important is a full P budget for the lake. In its 531 532 absence or where a sufficiently secure basic budget is not available, the difference between summer and winter lake water TP concentrations could indicate whether P release is 533 significant during the summer period. Jeppesen et al. (1997) collated TP results for 234 534 Danish lakes and their Fig. 4 shows that the summer concentration is more than twice the 535 winter value in shallow lakes that do not thermally stratify (mean depth <5 m) and have an 536 annual mean TP concentration greater than 100 μ g L⁻¹. The same behaviour was found by 537 Gibson et al. (1996) in 17 mostly Irish lakes; there is a mixture of shallow, non-stratifying 538 539 lakes and deeper, stratifying ones, but, based on the shallow lakes, the summer concentration is twice the winter value when the annual mean TP concentration is greater than 50 μ g L⁻¹. 540 541

542 Using the timescale results and this guidance on whether the internal P load is important, the
543 four lakes are assessed. Detailed lake water TP results from the Environmental Protection
544 Agency of Ireland's EDEN database were used to describe the seasonal changes.

The annual mean TP concentration in Lough Melvin varied from 19 to 29 μ g L⁻¹ between 546 1990 and 2015 (Table 1) and detailed results at six sites in the lake from 2007 to 2016 547 provide little evidence that summer release of P from the sediment is an important element 548 of the lake P budget (Fig. 2 in the Supplementary material). As the internal P load is 549 probably small, the timescale of reduction of sediment release estimated for the lake, a t₇₅ of 550 60 years, is not relevant to reduction in the water column TP concentration after external load 551 reduction. With a hydraulic residence time of 0.83 yr (Table 1), Melvin should respond 552 within a few years to changes in external P load. 553

554

The annual mean TP concentration in Lough Ramor varied from 39 to 100 μ g L⁻¹ between 2000 and 2015 and detailed results at five sites in the lake from 2007 to 2016 provide good evidence that summer release of P from the sediment is important, even with no results for January to March (Fig. 3 in the Supplementary material). As the internal P load is important, the timescale of reduction of sediment release largely determines the recovery of the lake to a concentration determined by the external P load, with an estimated t₇₅ of 33 years.

561

The annual mean TP concentration in Lough Sheelin varied from 13 to 57 μ g L⁻¹ between 562 1976 and 2015 and detailed results at five sites in the lake from 2007 to 2016 provide little 563 564 evidence that summer release of P from the sediment is an important element of the lake P budget (Fig. 4 in the Supplementary material). As the internal P load is probably small, the 565 estimated timescale of reduction of sediment release, a t₇₅ of 3 years, is not relevant to 566 reduction in the water column TP concentration after external load reduction. With a 567 hydraulic residence time of 0.55 yr (Table 1), Sheelin should respond within a few years to 568 changes in external P load. 569

571	A full P budget for 1975-1985 is available for Lough Neagh (Gibson et al. 1988) and the
572	description of sediment P release by Gibson et al. (2001) shows that it is the dominant
573	element; soluble P is fully recycled and particulate P sedimented. As internal load is
574	important, the timescale of reduction in sediment P release largely determines the long-term
575	recovery of the lake, with an estimated t75 of 41 years. Continued reduction in the inputs of
576	nitrogen as well phosphorus is needed before lower sediment P release influences
577	phytoplankton productivity.
578	
579	5 Conclusions
580	
581	Within the context of developing River Basin Management Plans, specifically the timescale
582	of achieving Good Status for lakes, the natural delay in recovery due to the internal P load in
583	some lakes needs to be considered.
584	
585	To support this analysis, a condensed model (Lewis/Penn) of the timescale of reduction of
586	labile P concentration in sediment that uses a diagenesis rate constant and burial rate constant
587	was applied to four lakes.
588	
589	The timescale (t_{75}) was 60 years in Lough Melvin, 3 in Ramor, 33 in Sheelin and 41 in
590	Neagh, although sediment P release is not important in Melvin and Sheelin.
591	
592	The diagenesis rate constants in the four lakes, combined with other values from the
593	literature, give mean and median (n=14) values of 0.0981 and 0.0426 yr ⁻¹ and so t_{75} values of
594	14 and 33 years; these overestimate the full timescales of reduction of labile P as burial is not
595	included.

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597	A review of other models indicates timescales of one to two decades and of lake monitoring
598	results also between one and two decades.
599	
600	It is desirable to estimate the timescale directly if sediment P release is important in a lake
601	and this model was found to be simple to apply. However, generally, the timescale should be
602	between one and three decades and this should be widely applicable.
603	
604	References
605	
606	Ahlgren, I. 1977 Role of sediments in the process of recovery of a eutrophicated lake. Golterman,
607	H.L. (ed), pp. 372-377, Pudoc, Amsterdam.
608	
609	Ahlgren, J., Transvik, L., Gogoll, A., Waldeback, M., Markides, K. and Rydin, E. 2005. Sediment
610	depth attenuation of biogenic phosphorus compounds measured by 31P nmr. Environmental Science
611	and Technology 39, 867-872.
612	
613	Andersen, J.M. 1976. An ignition method for determination of total phosphorus in lake sediments.
614	Water Research 10, 329-331.
615	
616	Appleby, P.G. 2001. Tracking Environmental Change Using Lake Sediments. Vol. 1: Basin
617	Analysis, Coring, and Chronological Techniques. Last, W.M. and Smol, J.P. (eds), pp. 171-203,
618	Kluwer Academic Publishers, Dordrecht.
619	

- Appleby, P.G. and Oldfield, F. 1978. The calculation of 210Pb dates assuming a constant rate of
 supply of unsupported 210Pb to the sediment. Catena 5, 1-8.
- 622
- Appleby, P.G., Nolan, P.J., Gifford, D.W., Godfrey, M.J., Oldfield, F., Anderson, N.J. and Battarbee,
- R.W. 1986. 210Pb dating by low background gamma counting. Hydrobiologia 141, 21-27.
- 625
- 626 Appleby, P.G., Richardson, N. and Nolan, P.J. 1992. Self-absorption corrections for well-type
- 627 germanium detectors. Nuclear Instruments and Methods in Physics Research Section B: Beam
- 628 Interactions with Materials and Atoms 71, 228-233.
- 629
- 630 Bengtsson, L. 1975. Phosphorus release from a highly eutrophic lake sediment. Verhandlungen
- 631 Internationale Vereinigung Limnologie fur theroetische und angewandte 19, 1107-1116.
- 632
- Bock, R. 1979. A Handbook of Decomposition Methods in Analytical Chemistry, International
 Textbook Company, Glasgow.
- 635
- Brett, M.T. and Benjamin , M.M. 2008. A review and reassessment of lake phosphorus retentionand the nutrient loading concept. Freshwater Biology 53, 194-211.
- 638
- Bunting, L., Leavitt, P.R., Gibson, C.E., McGee, E.J. and Hall, V.A. 2007. Degradation of water
- quality in Lough Neagh, Northern Ireland, by diffuse nitrogen flux from a phosphorus-rich
- 641 catchment. Limnology and Oceanography 53, 354-369.
- 642
- 643 Carey, C. and Rydin, E. 2011. Lake trophic status can be determined by the depth distribution of
- sediment phosphorus. Limnology and Oceanography 56, 2051-2063.

- Carignan, R. and Flett, R.J. 1982. Postdepositional mobility of phosphorus in lake sediments.
 Limnology and Oceanography 26, 361-366.
- 648
- 649 Carter, R.W.G. 1993. Lough Neagh. The Ecology of a Multipurpose Water Resource. Wood, R.B.

and Smith, R.V. (eds), pp. 35-57, Kluwer Academic Publishers, Dordrecht.

651

- Chapra, S. C. 1997. Surface Water-Quality Modeling. The McGraw-Hill Companies, New York.
- 654 Chapra, S.C. and Canale, R.P. 1991. Long-term phenomenological model of phosphorus and

oxygen for stratified lakes. Water Research 25, 707-715.

656

Cullen, P. and Forsberg, C. 1988. Experiences with reducing point sources of phosphorus to lakes.
Hydrobiologia 170, 321-336.

659

660 Dean, W.E. 1974. Determination of carbonate and organic matter in calcareous sediments and

sedimentary rocks by loss on ignition: comparison with other methods. Journal of Sedimentology 41,
242-248.

663

Dillon, P.J. and Rigler, F.H. 1975. A simple method for predicting the capacity of a lake for
development based on lake trophic status. Journal of the Fisheries Research Board of Canada 32,
1519-1531.

668	Ding, S., Xu, D., Bai, X., Yao, S., Fan, C. and Zhang, C. 2013. Speciation of organic phosphorus in
669	a sediment profile of Lake Taihu II. Molecular species and their depth attenuation. Journal of
670	Environmental Sciences 25, 925-932.

Dittrich, M., Wehrli, B. and Reichert, P. 2009. Lake sediments during the transient eutrophication

- period: Reactive-transport model and identifiability study. Ecological Modelling 220, 2751-2769.
- Eisenreich, S.J., Bannerman, R.T. and Armstrong, D.E. 1975. A simplified phosphorus analysis
 technique. Environmental Letters 9, 43-53.

677

Foy, R.H., Lennox, S.D. and Gibson, C.E. 2003. Changing perspectives on the importance of urban
phosphorus inputs as the cause of nutrient enrichment in Lough Neagh. The Science of the Total
Environment 310, 87-99.

681

Furumai, H. and Ohgaki, S. 1982. Fractional composition of phosphorus forms in sediments related
to release. Water Science and Technology 14, 215-226.

684

- Gibson, C.E., Foy, R.H. and Bailey-Watts, A.E. 1996. An analysis of the total phosphorus cycle in
 some temperate lakes, the response to enrichment. Freshwater Biology 35, 525-532.
- 687
- Gibson, C.E., Smith, R.V. and Stewart, D.A. 1988. A long term study of the phosphorus cycle in
- Lough Neagh, Northern Ireland. International Revue der gesamten Hydrobiologie 73, 249-257.

691	Gibson, C.E., Wang, G. and Foy, R.H. 2000. Silica and diatom growth in Lough Neagh: the
692	importance of internal recycling: the importance of internal recycling. Freshwater Biology 45, 285-
693	293.
694	
695	Gibson, C.E., Wang, G., Foy, R.H. and Lennox, S.D. 2001. The importance of catchment and lake
696	processes in the phosphorus budget of a large lake. Chemosphere 42, 215-220.
697	
698	Herodek, S. and Istvanovics, V. 1986. Mobility of phosphorus fractions in the sediments of Lake
699	Balaton. Hydrobiologia 135, 149-154.
700	
701	Hieljes, A.H.M. and Lijklema, L. 1980. Fractionation of inorganic phosphorus in calcareous lake
702	sediments. Journal of Environmental Quality 8, 130-132.

703

Hilton, J. 1985. A conceptual framework for predicting the occurrence of sediment focusing and 704 sediment redistribution in small lakes. Limnology and Oceanography 30, 1131-1143. 705

706

Hilton, J., Lishman, J.P. and Millington, A. 1986. A comparison of some rapid techniques for the 707 708 measurement of density in soft sediment. Sedimentology 33, 777-781.

709

- Horppila, J., Holmroos, H., Niemisto, J., Massa, J., Nygren, N., Schonach, P., Tapio, P. and 710
- Tammeorg, O. 2017. Variation of internal phosphorus loading and water quality in a 711
- hypereutrophic lake during 40 years of different management efforts. Ecological Engineering 103, 712 264-274. 713

715	Hupfer, M. and Lewandowski, J. 2008. Retention and early diagenetic transformation of
716	phosphorus in Lake Arendsee (Germany) - consequences for management strategies. Archiv fur
717	Hydrobiologie 164, 143-167.

Hupfer, M., Reitzel, K., Kleeberg, A. and Lewandowski, J. 2016. Long-term efficiency of lake
restoration by chemical phosphorus precipitation: Scenario analysis with a phosphorus balance

721 model. Water Research 97, 153-161.

722

Jensen, J.-P., Pedersen, A.R., Jeppesen, E. and Sondergaard, M. 2006. An empirical model

describing the seasonal dynamics of phosphorus in 16 shallow eutrophic lakes after external loading

reduction. Limnology and Oceanography 51, 791-800.

726

Jeppesen, E., Jensen, J.P., Sondergaard, M., Lauridsen, T., Pedersen, L.J. and Jensen, L. 1997. Topdown control if freshwater lakes: the role of nutrient state, submerged macrophytes and water depth.
Hydrobiologia 342/343, 151-164.

730

731 Jeppesen, E., M., Sondergaard, M., Jensen, J.-P., Havens, K.E., Anneville, O., Carvalho, L.,

732 Coveney, M.F., Deneke, R., Dokulil, M.T., Foy, B., Gerdeaux, D., Hampton, S.E., Hilt, S., Kangur,

- 733 K., Kohler, J., Lammens, E.H.H.R., Lauridsen, T.L., Manca, M., Miracle, M.R., Moss, B., Noges, P.,
- Persson, G., Phillips, G., Portielje, R., Romo, S., Schelske, C., Straile, D., Tatrai, I., Willen, E. and
- 735 Winder, M. 2005. Lake responses to reduced nutrient loading an analysis of contemporary long-
- term data from 35 case studies. Freshwater Biology 50, 1747-1771.

738	Jeppesen, E., Jensen, J.P., Sondergaard, M., Lauridsen, T., Pedersen, L.J. and Jensen, L. 1997. Top-
739	down control if freshwater lakes: the role of nutrient state, submerged macrophytes and water depth.
740	Hydrobiologia 342/343, 151-164.

Katsev, S.I., Tsandev, I., L'Heureux, I.L. and Rancourt, D.G. 2006. Factors controlling long-term
phosphorus efflux from lake sediments: Exploratory reactive-transport modeling. Chemical Geology
234, 127-147.

745

- Larsen, D.P., van Sickle, J., Malueg, K.W. and Smith, P.D. 1979. The effect of wastewater
- 747 phosphorus removal on Shagawa Lake, Minnesota: phosphorus supplies, lake phosphorus and

r48 chlorophyll a. Water Research 13, 1259-1272.

749

Lerman, A. 1979. Geochemical Processes: Water and Sediment Environments, Wiley-Interscience,Chichester.

752

Lewis, G.N., Auer, M.T., Xiang, X. and Penn, M.R. 2007. Modeling phosphorus flux in the

sediments of Onondaga Lake: Insights on the timing of lake response and recovery. Ecological

755 Modelling 209, 121-135.

756

- 757 Lukkari, K., Leivuori, M. and Hartikainen, H. 2007. Fractionation of sediment phosphorus
- revisited: II. Changes in phosphorus fractions during sampling and storage in the presence or absence

of oxygen. Limnology and Oceanography: Methods 5, 445-456.

760

761 Mackereth, F.J.H. 1969. A short core sampler for subaqueous deposits. Limnology and

762 Oceanography 14, 145-151.

105	7	υ	-
-----	---	---	---

764	May, L., Defew, L.H., Bennion, H. and Kirika, A. 2012. Historical changes (1905-2005) in external
765	phosphorus loads to Loch Leven, Scotland, UK. Hydrobiologia 681, 11-21.
766	
767	McCrackin, M.L., Jones, H.P., Jones, P.C. and Moreno-Mateos, D. 2016. Recovery of lakes and
768	coastal marine ecosystems from eutrophication: A global meta-analysis. Limnology and
769	Oceanography 62, 507-518.
770	
771	Moore, P.A., Reddy Jr, K.R. and Fisher, M.M. 1998. Phosphorus flux between sediment and
772	overlying water in Lake Okeechobee, Florida: spatial and temporal variations. Journal of
773	Environmental Quality 27, 1428-1439.
774	
775	Müller, B., Gachter, R. and Wuest, A. 2014. Accelerated water quality improvement during
776	oligotrophication in Peri-Alpine lakes. Environmental Science and Technology 48, 6671-6677.
777	
778	Murphy, J. and Riley, J.P. 1962. A modified single-solution method for the determination of
779	phosphate in natural waters. Analytica Chimica Acta 27, 31-36.
780	
781	Nembrini, G., Capobianco, J.A., Garcia, J. and Jacquet, JM. 1982. Interaction between interstitial
782	water and sediment in two cores of Lac Leman, Switzerland. Hydrobiologia 92, 363-375.
783	
784	Nürnberg, G.K. 1984. The prediction of internal phosphorus load in lakes with anoxic hypolimnia.
785	Limnology and Oceanography 29, 111-124.
786	

787	Nürnberg, G.K. 1996. Trophic state of clear and colored, soft- and hardwater lakes with special
788	consideration of nutrients, anoxia, phytoplankton and fish. Journal of Lake and Reservoir
789	Management 12, 432-447.
790	

Nürnberg, G.K. 1998. Prediction of annual and seasonal phosphorus concentrations in stratified and
polymictic lakes. Limnology and Oceanography 43, 1544-1552.

793

Ostrofsky, M.L. 1987. Phosphorus species in the surficial sediments of lakes of Eastern North
America. Canadian Journal of Fisheries and Aquatic Sciences 44, 960-966.

796

797 Ostrofsky, M.L. 2012. Determination of total phosphorus in lake sediments. Hydrobiologia 696,
798 199-203.

799

- Ozklundakci, D., Hamilton, D.P., McDowell, R. and Hill, S. 2014. Phosphorus dynamics in
- sediments of a eutrophic lake derived from 31P nuclear magnetic resonance spectroscopy. Marine
- and Freshwater Research 65, 70-80.

803

- Penn, M.R., Auer, M.T., van Orman, E. and Korienek, J.J. 1995. Phosphorus diagenesis in lake
 sediments: investigations using fractionation techniques. Marine and Freshwater Research 46, 89-99.
- Prairie, Y.T. 1989. Statistical models for the estimation of net phosphorus sedimentation in lakes.
 Aquatic Sciences 51, 192-210.

810	Psenner, R., Bostrom, B., Dinka, M., Pettersson, K. and Pukso, R. 1988. Fractionation of
811	phosphorus in suspended matter and sediment. Archiv fur Hydrobiologie Beiheft Ergebnisse der
812	Limnologie 30, 83-112.
813	
814	Quevauviller, P., Borchers, U., Thompson, C., and Simonart, T. 2008. The Water Framework
815	Directive. Ecological and Chemical Status Monitoring. John Wiley & Sons, Ltd, Chichester
816	
817	Reitzel, K., Ahlgren, J., Gogoll, A. and Rydin, E. 2006. Effects of aluminium treatment on
818	phosphorus carbon, and nitrogen distribution in lake sediment: A ³¹ P NMR study. Water Research
819	40, 647-654.
820	
821	Reitzel, K., Ahlgren, J., DeBrabandere, H., Waldeback, M., Gogoll, A., Transvik, L. and Rydin, E.
822	2007. Degradation rates of organic phosphorus in lake sediment. Biogeochemistry 82, 15-28.
823	
824	Rippey, B. and Anderson, N.J. 1996. Reconstruction of lake phosphorus loading and dynamics
825	using the sedimentary record. Environmental Science and Technology 30, 1786-1788.
826	
827	Ruban, V., Lopez-Sanchez, J.F., Pardo, P., Rauret, G., Muntau, H. and Ouevauviller, P. 1999.
828	Selection and evaluation of sequential extraction procedures for the determination of phosphorus
829	forms in lake sediment. Journal of Environmental Monitoring 1, 51-56.
830	
831	Schippers, P., van de Weerd, H., J., d.K., de Jong, B. and Scheffer, M. 2006. Impacts of agricultural
832	phosphorus use in catchments on shallow lake water quality: About buffers, time delays and
833	equilibria. Science of the Total Environment 369, 280-294.
834	

835	Shatwell, T. and Kohler, J. 2019. Decreased nitrogen loading controls summer cyanobacterial
836	blooms without promoting nitrogen-fixing taxa, Long-term response of a shallow lake. Limnology
837	and Oceanography 64, 5166-5178.
838	
839	Sondergaard, M., Windolf, J. and Jeppensen, E. 1996. Phosphorus factions and profiles in the
840	sediment of shallow Danish lakes as related to phosphorus load, sediment composition and lake
841	chemistry. Water Research 30, 92-1002.
842	
843	Spears, B.M., Carvalho, L., Perkins, R., Kirika, A. and Paterson, D.M. 2012. Long-term variation
844	and regulation of internal phosphorus loading in Loch Leven. Hydrobiologia 681, 23-33.
845	
846	Stronge, K.M., Smith, R.V. and Lennox, S. 1998. Predicting the spring algal biomass in Lough
847	Neagh using time series analysis. Freshwater Biology 39, 593-600.
848	
849	Van der Molen, D.T. 1991. A simple dynamic model for the simulation of the release of phosphorus
850	from sediments in shallow eutrophic systems. Water Research 25, 737-744.
851	
852	Verdonschot, P.F.M., Spears, B.M., Feld, C.K., Brucet, S., Keizer-Velk, H., Borja, A., Elliott, M.,
853	Kernan, M. and Johnson, R.K. 2013. A comparative review of recovery processes in rivers, lakes,
854	estuarine and coastal waters. Hydrobiologia 704, 453-474.
855	
856	Wilson, T.A., Amirbahman, A., Norton, S.A. and Voytek, M.A. 2010. A record of phosphorus
857	dynamics in oligotrophic lake sediment. Journal of Paleolimnology 44, 279-294.
858	

Wood, R.B. and Smith, R.V. 1993. Lough Neagh. The Ecology of a Multipurpose Water Resource,
Kluwer Academic Publishers, Dordrecht.

861

862 Zar, J. H. 2010. Biostatistical Analysis, Pearson Prentice Hall, Ney Jersey.

863

864 Acknowledgements

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We would like to thank the two funders, the Environmental Protection Agency of Ireland for
the work on Lough Melvin, Ramor and Sheelin through the DETECT Project (2015-W-LS-9)
and the Department for Agriculture, Environment and Rural Affairs in Northern Ireland for
Lough Neagh.

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We would also like to thank the following for the use of boats and invaluable assistance to
retrieve sediment cores from the lakes: Hugh Gillespie and Kevin McCloskey for Lough
Melvin, Fergus Lynch for Lough Ramor and Stephen Ryan, Frankie Conlon and Martin
Devlin for Lough Neagh. Thanks go to Handong Yang of the Environmental Radiometric
Facility at University College London for completing the dating of the sediment cores.
We would also like to thank Dr Shane O'Boyle of the Environmental Protection Agency for
valuable comments on the manuscript and two anonymous reviewers for comments and

879 suggestions that improved the clarity of the text.

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884	Figure	captions
001	I Igui v	cuptions

Fig. 1. The Lewis/Penn model of the rate of change of labile P in lake sediment. The slowdiagenesis rate constant only is used as the focus in the long-term timescale.

Fig. 2. Variation of Fraction A and labile P with depth in sediment cores from Lough Melvin
(MEL5), Ramor (RAM5) and Sheelin (SHE5).

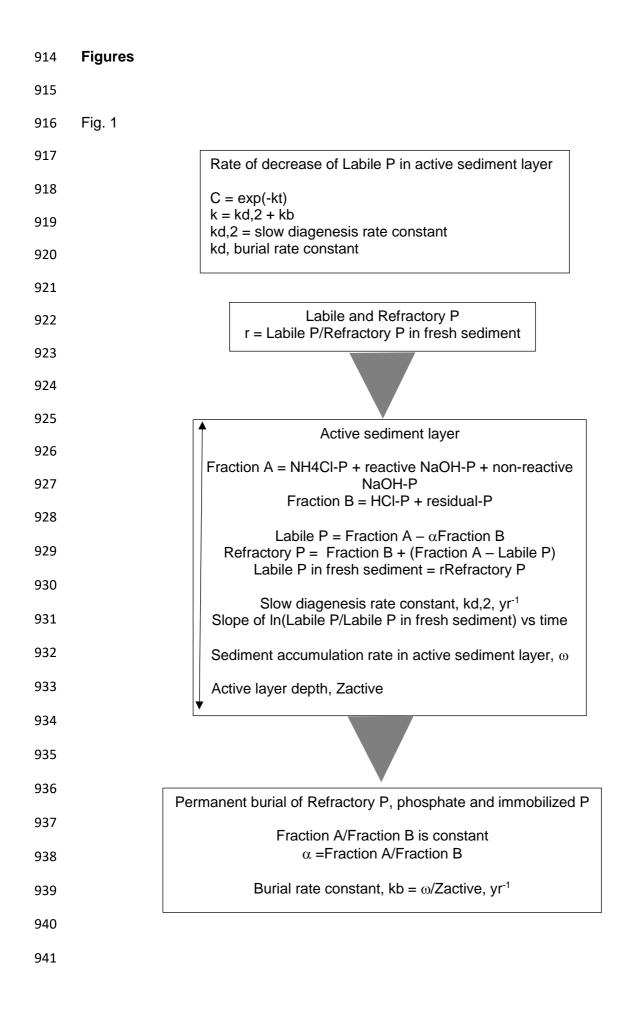
Fig. 3. Variation of Fraction A and labile P with depth in sediment cores from Lough Neagh,
LN11, LN15, LN17, LN18 and LN19.

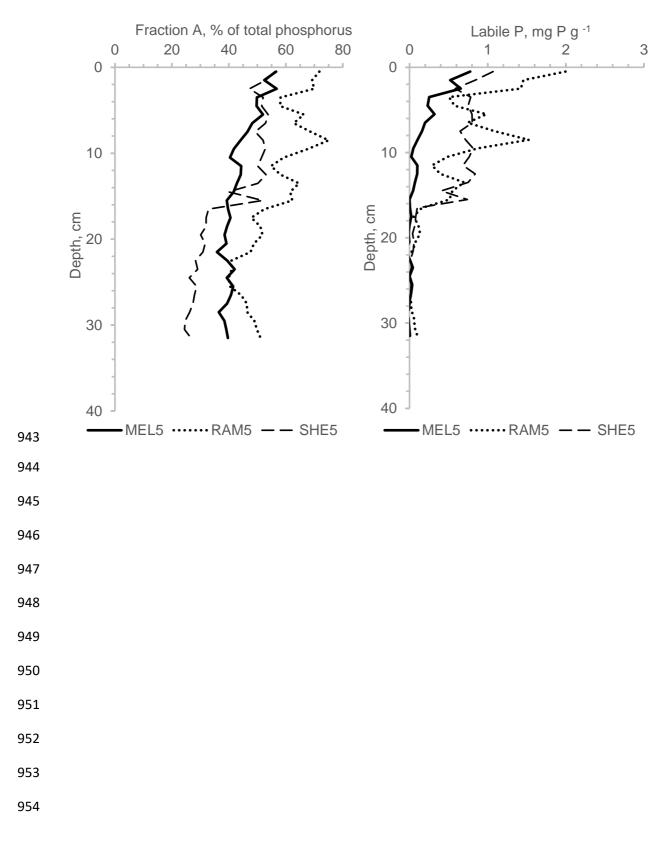
Fig. 4. Variation of the natural logarithm of the ratio of labile P to labile P in fresh sediment

896 with date in sediment cores from Lough Melvin (MEL5), Ramor (RAM5) and Sheelin

897 (SHE5). The linear regressions are shown and the open circle data points were omitted.

Fig. 5. Variation of the natural logarithm of the ratio of labile P to labile P in fresh sediment
with date in sediment cores from Lough Neagh, LN11, LN15, LN17, LN18 and LN19. The
linear regressions are shown and the open circle data points were omitted.





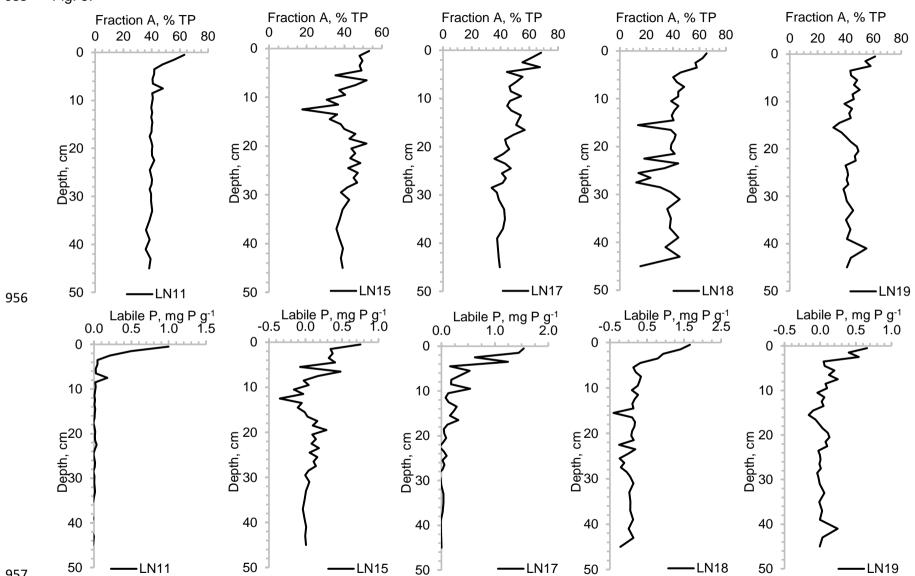
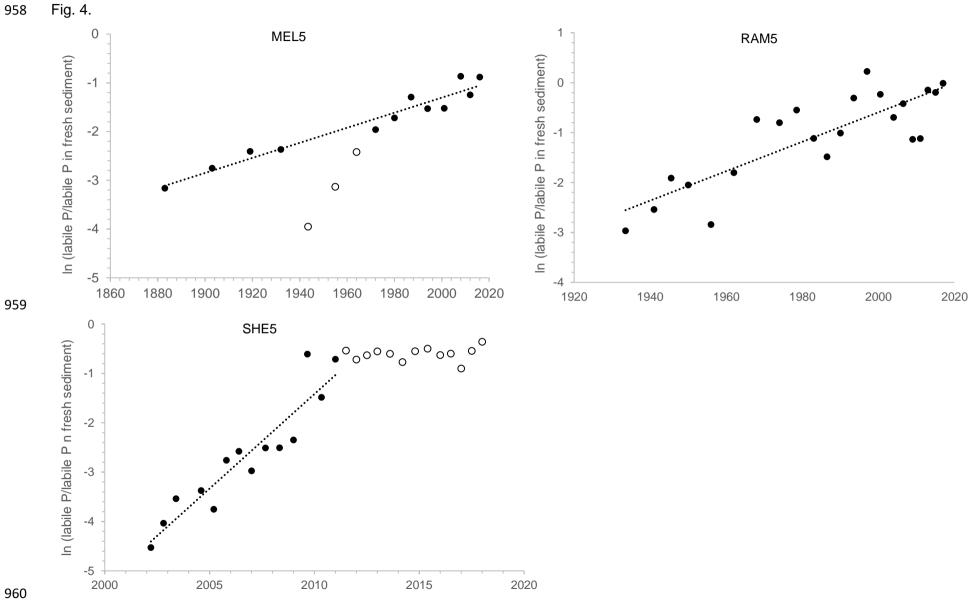
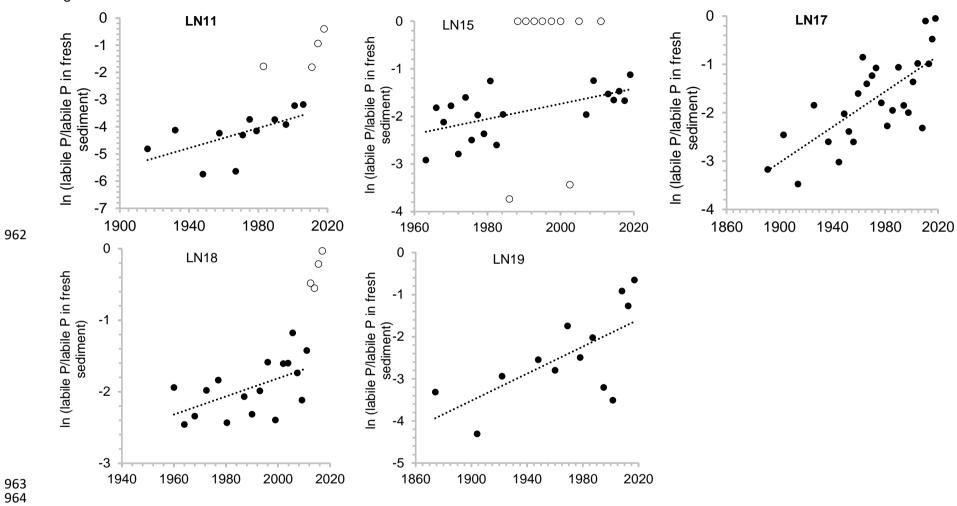


Fig. 3.





961 Fig. 5.

969 Tables

Table 1. The basic limnological characteristics of the lakes. The range of annual mean total phosphorus (TP) concentration and the time period and the sediment type, water depth at the sediment coring site, date of coring and sediment core designation are also given.

Lough	Latitude/longitude	Area,	Mean	Max.	Hydraulic	Alkalinity,	Annual	Sediment	Water depth at
		km²	depth, m	depth, m	residence	meq L ⁻¹	mean TP,	type	coring site, m,
					time, yr		μg L⁻¹		date cored, core
									designation
Melvin	54.5933/6.4162	22	10.9	45	0.83	1.13	19-29 (1990-2015)	Siliceous	12.6, 5 Oct 2017, MEL5
Ramor	53.8144/7.0642	7.12	3.0	5.5	0.17	1.14	39-100 (2000-2015)	Siliceous	3.4, southern basin, 8 Nov. 2017, RAM5
Sheelin	53.8125/7.3213	4.4	4.5	15	0.55	3.12	13-57 (1976-2015)	Calcareous, 43.3 % DS CaCO3	8.7, 16 May 2018, SHE5
Neagh	54.5933/6.4162	383	8.9	34	1.2	2.16	106-141 (2007-2016)	Siliceous	10.3-12.9, 4 Apr 2019 LN11, 19 Feb 2019 LN15, 4 Dec 2018 LN17, 4 Dec 2018 LN18 & 19 Mar 2019 LN19

Table 2. The measured values of sediment accumulation rate (ω), active sediment layer depth (Zactive), slow diagenesis rate constant (kd,2, \pm SE), burial rate constant (kb) and overall rate constant (k = kd,2+kb) in a sediment core from each of Lough Melvin, Ramor and Sheelin and

Lough,	ω, cm yr⁻¹	Zactive,	kd,2, yr⁻¹	kb, yr⁻¹	k, yr⁻¹	kb/k, %	t ₅₀ , yr	t ₇₅ , yr	t ₉₀ , yr
core		cm							
Melvin, MEL5	0.113	15	0.0155 (0.00123)	0.00753	0.0230	32.8	30	60	100
Ramor, RAM5	0.265	22	0.0295 (0.00436)	0.0120	0.0415	29.0	17	33	56
Sheelin, SHE5	1.75	28	0.383 (0.0431)	0.0625	0.446	14.0	1.6	3.1	5.2
Neagh, LN11	0.229	8	0.0184 (0.00730)	0.0286	0.0470	60.9	15	30	49
Neagh, LN15	0.519	29	0.0159 (0.00490)	0.0179	0.0338	52.9	21	41	68
Neagh, LN17	0.304	21	0.0184 (0.00344)	0.0145	0.0329	44.0	21	42	70
Neagh, LN18	0.377	23	0.0138 (0.00538)	0.0164	0.0302	54.3	23	46	76
Neagh, LN19	0.145	10	0.0161 (0.00545)	0.0145	0.0306	47.4	23	45	75

984 five cores from Lough Neagh. The burial to total loss of labile P from the active layer (kb/k) and times to reduce labile P by 50, 75 and 90 % are 985 also given.

Table 3. Depth of the layer of diagenesis of labile P in lake sediment and the evidence on which it is based.

Lake	Depth, cm	Evidence	Reference

Lake Memphremagog	15-20	Variation of total, mobile and	Carignan & Flett (1982)
		interstitial phosphorus	
		concentrations in the sediment	
Lac Leman	20	Variation of iron, manganese and	Nembrini et al. (1982)
		soluble phosphorus in interstitial	
		water	
Nine Danish lakes	10-30	Change in P fraction concentrations	Sondergaard et al. (1996)
		with depth in the sediment	
Lake Okeechobee	10-25	Variation of soluble P concentration	Moore et al. (1998)
		with depth in interstitial water	
Lake Okaro	At least 10	Change in concentration with depth	Ozklundakci et al. (2014)
		of orthophosphate monoesters,	
		orthophosphate diesters,	
		pyrophosphate and polyphosphate	
		in sediment	
Eight lakes	3-14	Depth to background total	Hupfer et al. (2016)
		phosphorus concentration in	
		sediment core	

Table 4. Values of the slow diagenesis rate constant for P in lake sediment. The timescale that the value applies to is noted. Where given, the average of values for phospholipids and DNA, and for teichoic acids and DNA, were averaged and presented as phosphate diesters. Note that Lake Erken has three entries in the table. Methodology: 1 sedimentary P fractions in dated sediment core; 2 calibration of model; 3 change of 1012 concentration (estimated by P-31 NMR) with depth in dated core. The trophic state is given as annual mean TP concentration (μg L⁻¹) or as a
 1013 trophic class.

Lake	Rate constant, yr ⁻¹	Timescale	Methodology	Trophic state	Reference
Whole sediment in Lough Melvin	0.0115	133 yr	1	19-29	This work
Whole sediment in Lough Ramor	0.0295	83 yr	1	39-100	This work
Whole sediment in Lough Sheelin	0.383	9 yr	1	13-57	This work
Whole sediment in Lough Neagh	0.0165	93 yr	1	106-141	This work
Whole sediment in Lake Onondaga	0.11	22 yr	1	20-70	Penn et al. (1995
Mud Pond	0.0114	100 yr	1	3	Wilson et al. (2010)
Little Long Pond	0.00701	100 yr	1	6	Wilson et al. (2010)
Upper Haddock Pond	0.0291	50 yr	1	7	Wilson et al. (2010)
Furesco, Esrom and Glumso	0.07	Four mths	2 using laboratory measured anaerobic release rate from three eutrophic lakes	All eutrophic	Cited in Penn et al. (1995)
White	0.0073 in lower sediment	Seasonal	2 for a eutrophic lake	30	Cited in Penn et al. (1995)
Esrom	0.18	Seasonal	2 using field measurements in the eutrophic lake	Eutrophic	Cited in Penn et al. (1995)
Lake Warner	0.20 (0.37)	Seasonal	2 using field measurements in the eutrophic lake, although the authors used a higher value (0.37) to illustrate delayed recovery	90	Cited in Penn et al. (1995)

Orthophosphate monoesters, orthophosphate diesters, pyrophosphates in Lake Erken	0.0387 (0.030, 0.033, omitting 0.053 due to short timescale)	100 yr	3	27	Ahlgren et al. (2005)
Orthophosphate monoesters, phospholipids, DNA and polyphosphates in Lake Sonderby	0.173 (0.161, 0.184, omitting 0.866 due to short timescale)	Decades	3	1500	Reitzel et al. (2006)
Orthophosphate monoesters, orthophosphate diesters and polyphosphates in humics in Lake Erken	0.0465 (0.00797,0.0155,0.116)	80 yr	3	27	Reitzel et al. (2007)
Orthophosphate monoesters, orthophosphate diesters and polyphosphates in non-humics in lake Erken	0.103 (0.0118,0.0315,0.347)	80 yr	3	27	Reitzel et al. (2007)
Orthophosphate monoesters, phospholipids, DNA and pyrophosphates in Lake Taihu (Meilang Bay)	0.103 (0.0257, 0.0537, 0.231)	12 yr	3	90	Ding et al. (2013)
Orthophosphate monoesters,	0.0525 (0.030, 0.035, 0.058, 0.087)	40 yr	3	20-220	Ozklundakci et al. (2014)

orthophosphate			
diesters,			
pyrophosphates,			
polyphosphonates			
in Lake Okaro			

1016 Supplementary material

1017

1018Table 1. Sediment phosphorus fraction concentrations (mg P g ¹ DS) in sediment core1019MEL5 in Lough Melvin.

Depth, cm	NH4CI-P	Reactive NaOH-P	Non-reactive NaOH-P	HCI-P	Residual-P	Total P
0.5	0.002	1.133	0.409	0.303	0.883	2.731
1.5	0.002	1.191	0.072	0.392	0.756	2.413
2.5	0.001	1.071	0.223	0.322	0.664	2.280
3.5	0.000	0.484	0.243	0.235	0.499	1.461
4.5	0.000	0.451	0.220	0.238	0.441	1.350
5.5	0.000	0.624	0.174	0.286	0.452	1.537
6.5	0.000	0.472	0.181	0.261	0.441	1.356
7.5	0.000	0.418	0.209	0.227	0.492	1.346
8.5	0.000	0.365	0.207	0.237	0.488	1.296
9.5	0.000	0.350	0.164	0.234	0.484	1.233
10.5	0.000	0.319	0.170	0.238	0.483	1.209
11.5	-0.001	0.365	0.171	0.230	0.442	1.207
12.5	0.000	0.359	0.184	0.239	0.449	1.231
13.5	-0.001	0.344	0.179	0.241	0.457	1.220
14.5	0.000	0.312	0.169	0.222	0.451	1.153
15.5	0.000	0.282	0.141	0.244	0.410	1.077
16.5	0.000	0.275	0.133	0.210	0.409	1.025
17.5	0.000	0.291	0.141	0.246	0.388	1.066
18.5	0.000	0.265	0.126	0.254	0.349	0.994
19.5	0.000	0.247	0.130	0.215	0.387	0.978
20.5	0.000	0.283	0.136	0.245	0.404	1.067
21.5	0.000	0.234	0.102	0.255	0.346	0.937
22.5	0.000	0.271	0.124	0.236	0.369	1.000
23.5	-0.001	0.280	0.132	0.203	0.364	0.979
24.5	0.000	0.245	0.100	0.213	0.320	0.877
25.5	0.000	0.277	0.116	0.247	0.305	0.945
26.5	0.000	0.249	0.110	0.198	0.323	0.880
27.5	-0.001	0.243	0.102	0.212	0.319	0.876
28.5	0.000	0.227	0.083	0.271	0.269	0.849
29.5	-0.001	0.247	0.097	0.215	0.335	0.894
30.5	-0.001	0.251	0.103	0.202	0.349	0.904
31.5	-0.001	0.253	0.101	0.190	0.346	0.889

1021 Table 2. Sediment phosphorus fraction concentrations (mg P g⁻¹ DS) in sediment core 1022 RAM5 in Lough Ramor.

Depth,	NH4CI-P	Reactive	Non-reactive	HCI-P	Residual-P	Total P
cm		NaOH-P	NaOH-P			
0.5	0.011	1.249	1.682	0.508	0.651	4.101
1.5	0.009	0.953	1.324	0.463	0.553	3.302
2.5	0.011	1.130	1.055	0.426	0.519	3.142
3.5	0.010	1.611	-0.392	0.415	0.471	2.114
4.5	0.011	1.463	-0.059	0.539	0.487	2.440
5.5	0.011	1.988	-0.317	0.485	0.370	2.537
6.5	0.006	2.133	-0.695	0.410	0.446	2.301
7.5	0.003	1.409	0.321	0.355	0.433	2.522
8.5	0.003	1.031	1.064	0.312	0.386	2.797
9.5	0.003	0.774	0.660	0.318	0.366	2.119
10.5	0.006	1.195	-0.085	0.377	0.391	1.884
11.5	0.004	1.174	-0.296	0.336	0.392	1.610

12.5	0.003	0.914	0.044	0.330	0.362	1.653
13.5	0.003	0.875	0.385	0.284	0.407	1.954
14.5	0.002	0.603	0.517	0.279	0.422	1.823
15.5	0.002	0.391	0.618	0.296	0.316	1.623
16.5	0.001	0.405	0.190	0.281	0.258	1.135
17.5	0.001	0.293	0.160	0.235	0.259	0.948
18.5	0.001	0.355	0.163	0.225	0.274	1.018
19.5	0.001	0.396	0.166	0.211	0.313	1.087
20.5	0.001	0.377	0.148	0.227	0.324	1.078
21.5	0.000	0.269	0.136	0.205	0.242	0.852
22.5	0.001	0.215	0.126	0.182	0.313	0.836
23.5	0.001	0.201	0.125	0.175	0.296	0.797
24.5	0.000	0.198	0.129	0.186	0.305	0.818
25.5	0.000	0.187	0.120	0.162	0.291	0.760
26.5	0.000	0.217	0.135	0.193	0.250	0.796
27.5	0.000	0.225	0.138	0.177	0.244	0.785
28.5	0.000	0.253	0.149	0.212	0.252	0.866
29.5	0.000	0.260	0.143	0.196	0.226	0.827
30.5	0.000	0.231	0.140	0.174	0.200	0.746
31.5	0.000	0.301	0.155	0.217	0.219	0.892

1024Table 3. Sediment phosphorus fraction concentrations (mg P g⁻¹ DS) in sediment core1025SHE5 in Lough Sheelin.

Depth,	NH4CI-P	Reactive NaOH-P	Non-reactive NaOH-P	HCI-P	Residual-P	Total P
cm						
0.5	0.047	1.061	0.386	0.542	0.625	2.661
1.5	0.031	0.891	0.323	0.523	0.583	2.352
2.5	0.031	0.939	0.011	0.558	0.538	2.078
3.5	0.029	0.900	0.250	0.538	0.549	2.267
4.5	0.028	0.886	0.226	0.559	0.513	2.212
5.5	0.027	0.793	0.343	0.521	0.480	2.164
6.5	0.028	0.851	0.309	0.562	0.497	2.247
7.5	0.030	0.859	0.147	0.568	0.499	2.103
8.5	0.026	0.769	0.286	0.527	0.470	2.078
9.5	0.028	0.824	0.355	0.600	0.480	2.287
10.5	0.029	0.891	0.242	0.695	0.401	2.258
11.5	0.030	0.933	0.111	0.613	0.457	2.143
12.5	0.033	1.049	0.157	0.655	0.443	2.337
13.5	0.033	1.080	0.060	0.708	0.455	2.337
14.5	0.031	1.199	-0.326	0.903	0.464	2.271
15.5	0.027	0.746	0.345	0.584	0.454	2.155
16.5	0.022	0.235	0.126	0.374	0.408	1.165
17.5	0.022	0.204	0.130	0.348	0.408	1.111
18.5	0.021	0.164	0.124	0.300	0.358	0.967
19.5	0.017	0.126	0.109	0.258	0.325	0.834
20.5	0.017	0.134	0.117	0.265	0.312	0.844
21.5	0.016	0.119	0.113	0.243	0.314	0.806
22.5	0.012	0.099	0.089	0.216	0.291	0.707
23.5	0.014	0.115	0.098	0.242	0.313	0.782
24.5	0.011	0.098	0.082	0.223	0.319	0.733
25.5	0.013	0.109	0.098	0.235	0.312	0.766
26.5	0.013	0.095	0.087	0.216	0.291	0.702
27.5	0.013	0.100	0.083	0.212	0.305	0.713
28.5	0.013	0.089	0.083	0.205	0.314	0.704
29.5	0.010	0.088	0.076	0.198	0.336	0.709
30.5	0.010	0.071	0.073	0.176	0.300	0.630

31.5 0.009 0.080 0.078	0.205	0.256	0.628
------------------------	-------	-------	-------

HCI-P

0.490

0.448

0.372

0.296

0.283

0.327

0.296

Residual-P

0.445

0.346

0.421

0.339

0.365

0.239

0.238

Total P

2.536

1.804

1.515

1.093

1.113

0.954

0.903

1.256

0.850

0.822

0.868

0.819

0.856

0.836

0.801

0.745

0.753

0.821

0.757

0.760

0.750

0.787

0.839

0.829

0.768

0.755

0.816

0.807

0.791

0.824

0.792

0.785

0.781

0.812

0.783

0.805

0.800

0.809

1026 Table 4. Sediment phosphorus fraction concentrations (mg P g⁻¹ DS) in sediment core LN11 1027 1028 in Lough Neagh.

Non-reactive

NaOH-P

0.605

0.311

0.292

0.248

0.217

0.192

0.156

Depth,

cm

0.5

1.5 2.5

3.5

4.5

5.5

6.5

NH4CI-P

0.056

0.005

0.002

0.001

0.000

0.001

0.001

Reactive

NaOH-P

0.941

0.694

0.429

0.209

0.248

0.195

0.212

0.432 0.374 0.279 7.5 0.002 0.168 8.5 0.000 0.168 0.176 0.251 0.255 9.5 0.001 0.176 0.159 0.226 0.261 10.5 0.177 0.258 0.260 0.001 0.172 11.5 0.001 0.161 0.164 0.242 0.252 12.5 0.001 0.169 0.175 0.258 0.252 13.5 0.156 0.175 0.245 0.260 0.002 14.5 0.001 0.156 0.167 0.228 0.250 15.5 0.001 0.146 0.151 0.234 0.213 16.5 0.001 0.151 0.148 0.241 0.212 0.227 17.5 0.171 0.145 0.277 0.000 18.5 0.244 0.211 0.001 0.156 0.143 19.5 0.001 0.133 0.174 0.243 0.209 20.5 0.001 0.127 0.174 0.228 0.221 21.5 0.001 0.146 0.171 0.241 0.229 22.5 0.001 0.164 0.186 0.251 0.236 0.230 23.5 0.001 0.164 0.169 0.265 0.232 24.5 0.001 0.134 0.162 0.239 25.5 0.135 0.213 0.244 0.001 0.162 26.5 0.001 0.133 0.194 0.255 0.232 27.5 0.000 0.151 0.171 0.252 0.233 0.244 28.5 0.001 0.155 0.150 0.243 29.5 0.001 0.170 0.157 0.265 0.232 0.001 0.159 0.153 0.235 0.244 31.0 0.176 33.0 0.001 0.141 0.237 0.232 35.0 0.142 0.228 0.000 0.158 0.252 0.126 37.0 0.000 0.165 0.273 0.248 39.0 0.000 0.168 0.134 0.251 0.230 41.0 0.000 0.154 0.134 0.251 0.266 43.0 0.000 0.148 0.251 0.235 0.165 45.0 0.001 0.168 0.140 0.262 0.237 Table 5. Sediment phosphorus fraction concentrations (mg P g⁻¹ DS) in sediment core LN15 in Lough Neagh.

1031

1029

Depth, cm	NH4CI-P	Reactive NaOH-P	Non-reactive NaOH-P	HCI-P	Residual-P	Total P
0.5	0.027	1.132	0.524	0.620	0.862	3.165
1.5	0.008	0.713	0.349	0.500	0.661	2.232
2.5	0.006	0.698	0.318	0.476	0.560	2.058
3.5	0.004	0.599	0.384	0.454	0.611	2.052
4.5	0.002	0.806	0.352	0.561	0.640	2.361
5.5	0.000	0.213	0.260	0.363	0.508	1.345

6.5	0.006	0.884	0.261	0.588	0.481	2.220
7.5	0.002	0.437	0.201	0.338	0.418	1.398
8.5	0.000	0.167	0.288	0.338	0.430	1.223
9.5	0.001	0.358	0.224	0.390	0.471	1.445
10.5	0.000	0.177	0.202	0.357	0.503	1.239
11.5	0.000	0.150	0.215	0.273	0.357	0.995
12.5	0.001	0.108	0.074	0.375	0.476	1.033
13.5	0.001	0.311	0.149	0.410	0.403	1.273
14.5	0.000	0.147	0.182	0.330	0.364	1.023
15.5	0.000	0.229	0.199	0.334	0.364	1.126
16.5	0.001	0.280	0.176	0.360	0.326	1.143
17.5	0.001	0.400	0.225	0.389	0.348	1.363
18.5	0.001	0.484	0.150	0.454	0.401	1.490
19.5	0.001	0.468	0.224	0.310	0.336	1.339
20.5	0.001	0.260	0.224	0.272	0.354	1.110
21.5	0.001	0.349	0.201	0.301	0.351	1.202
22.5	0.001	0.282	0.191	0.283	0.344	1.101
23.5	0.001	0.310	0.244	0.256	0.333	1.143
24.5	0.001	0.212	0.198	0.245	0.322	0.977
25.5	0.001	0.351	0.193	0.271	0.340	1.156
26.5	0.000	0.286	0.191	0.265	0.320	1.062
27.5	0.001	0.247	0.245	0.253	0.305	1.051
28.5	0.001	0.153	0.200	0.220	0.277	0.850
29.5	0.000	0.128	0.180	0.226	0.273	0.807
31.0	0.001	0.136	0.206	0.192	0.270	0.804
33.0	0.000	0.126	0.183	0.225	0.257	0.791
35.0	0.000	0.127	0.158	0.232	0.241	0.758
37.0	0.000	0.137	0.148	0.238	0.274	0.796
39.0	0.000	0.130	0.164	0.220	0.273	0.787
41.0	0.000	0.132	0.170	0.205	0.262	0.770
43.0	0.000	0.125	0.140	0.188	0.242	0.695
45.0	0.000	0.135	0.164	0.197	0.268	0.765

Table 6. Sediment phosphorus fraction concentrations (mg P g⁻¹ DS) in sediment core LN17 in Lough Neagh. 1034

Depth,	NH4CI-P	Reactive	Non-reactive	HCI-P	Residual-P	Total P
cm		NaOH-P	NaOH-P			
0.5	0.039	1.507	0.652	0.522	0.502	3.221
1.5	0.053	1.788	0.552	0.857	0.616	3.865
2.5	0.025	0.818	0.464	0.494	0.567	2.367
3.5	0.021	1.506	0.281	0.486	0.389	2.682
4.5	0.012	0.512	0.301	0.435	0.597	1.857
5.5	0.009	0.790	0.309	0.387	0.509	2.004
6.5	0.002	0.610	0.309	0.399	0.481	1.801
7.5	0.001	0.490	0.252	0.421	0.446	1.610
8.5	0.000	0.368	0.270	0.308	0.407	1.353
9.5	0.002	0.971	0.198	0.465	0.519	2.155
10.5	0.000	0.286	0.209	0.273	0.298	1.067
11.5	0.000	0.209	0.177	0.241	0.239	0.866
12.5	0.000	0.269	0.197	0.263	0.252	0.980
13.5	0.000	0.442	0.184	0.276	0.254	1.157
14.5	0.000	0.378	0.176	0.254	0.248	1.055
15.5	0.000	0.252	0.172	0.197	0.214	0.835
16.5	0.003	0.472	0.143	0.242	0.228	1.088
17.5	0.003	0.218	0.125	0.172	0.187	0.706
18.5	0.000	0.161	0.136	0.196	0.194	0.687

19.5	0.001	0.196	0.101	0.194	0.183	0.675
20.5	0.001	0.302	0.057	0.199	0.223	0.781
21.5	0.001	0.162	0.126	0.189	0.211	0.689
22.5	0.001	0.137	0.105	0.220	0.217	0.680
23.5	0.001	0.177	0.109	0.177	0.200	0.664
24.5	0.001	0.238	0.111	0.196	0.196	0.740
25.5	0.001	0.159	0.121	0.218	0.188	0.687
26.5	0.001	0.214	0.112	0.229	0.191	0.746
27.5	0.001	0.131	0.125	0.163	0.199	0.620
28.5	0.001	0.106	0.099	0.193	0.211	0.610
29.5	0.001	0.120	0.115	0.183	0.209	0.628
31.0	0.001	0.119	0.120	0.181	0.197	0.618
33.0	0.001	0.230	0.062	0.195	0.202	0.690
35.0	0.001	0.112	0.144	0.155	0.185	0.597
37.0	0.001	0.185	0.051	0.158	0.173	0.568
39.0	0.001	0.095	0.113	0.181	0.166	0.555
41.0	0.000	0.076	0.112	0.159	0.146	0.493
43.0	0.001	0.088	0.137	0.177	0.180	0.582
45.0	0.001	0.076	0.152	0.170	0.180	0.579

Table 7. Sediment phosphorus fraction concentrations (mg P g⁻¹ DS) in sediment core LN18 in Lough Neagh. 1037

Depth, cm	NH4CI-P	Reactive NaOH-P	Non-reactive NaOH-P	HCI-P	Residual-P	Total P
0.5	0.085	1.745	0.448	0.634	0.570	3.482
1.5	0.041	1.558	0.439	0.643	0.586	3.268
2.5	0.017	1.150	0.365	0.619	0.538	2.689
3.5	0.006	0.674	0.578	0.404	0.504	2.166
4.5	0.003	0.539	0.276	0.472	0.496	1.786
5.5	0.000	0.266	0.299	0.351	0.490	1.406
6.5	0.000	0.339	0.290	0.391	0.445	1.464
7.5	0.000	0.349	0.395	0.362	0.428	1.534
8.5	0.002	0.619	0.242	0.532	0.562	1.957
9.5	0.001	0.438	0.286	0.429	0.492	1.646
10.5	0.001	0.305	0.208	0.352	0.466	1.332
11.5	0.001	0.570	0.175	0.479	0.462	1.687
12.5	0.000	0.452	0.193	0.453	0.475	1.573
13.5	0.000	0.342	0.195	0.417	0.422	1.377
14.5	0.004	0.380	0.182	0.409	0.423	1.398
15.5	0.000	0.097	0.080	0.554	0.574	1.305
16.5	0.000	0.427	0.158	0.479	0.457	1.521
17.5	0.000	0.438	0.175	0.381	0.462	1.456
18.5	0.005	0.405	0.203	0.408	0.471	1.492
19.5	0.000	0.257	0.263	0.411	0.405	1.336
20.5	0.000	0.278	0.155	0.345	0.352	1.129
21.5	-0.001	0.317	0.211	0.362	0.386	1.276
22.5	0.000	0.120	0.075	0.478	0.389	1.062
23.5	0.000	0.391	0.147	0.352	0.332	1.222
24.5	0.000	0.176	0.124	0.274	0.319	0.893
25.5	-0.001	0.059	0.052	0.294	0.383	0.788
26.5	-0.001	0.082	0.075	0.252	0.266	0.675
27.5	0.000	0.046	0.031	0.272	0.274	0.623
28.5	0.000	0.131	0.127	0.232	0.359	0.850
29.5	-0.001	0.125	0.152	0.217	0.229	0.723
31.0	-0.001	0.166	0.196	0.215	0.223	0.800
33.0	-0.001	0.154	0.129	0.269	0.238	0.789

35.0	-0.001	0.146	0.160	0.222	0.267	0.795
37.0	-0.001	0.204	0.142	0.287	0.287	0.920
39.0	-0.001	0.188	0.178	0.224	0.238	0.827
41.0	-0.001	0.140	0.139	0.254	0.276	0.809
43.0	-0.001	0.185	0.183	0.217	0.230	0.813
45.0	-0.001	0.170	-0.049	0.338	0.319	0.778

Table 8. Sediment phosphorus fraction concentrations (mg P g⁻¹ DS) in sediment core LN19 in Lough Neagh.

Depth, cm	NH4CI-P	Reactive NaOH-P	Non-reactive NaOH-P	HCI-P	Residual-P	Total P
-						1
0.5	0.012	0.778	0.399	0.323	0.429	1.941
1.5	0.003	0.593	0.397	0.378	0.461	1.832
2.5	0.002	0.619	0.489	0.356	0.446	1.911
3.5	0.000	0.263	0.343	0.310	0.478	1.394
4.5	0.001	0.324	0.314	0.344	0.467	1.450
5.5	0.004	0.509	0.302	0.406	0.461	1.682
6.5	0.001	0.460	0.239	0.361	0.457	1.517
7.5	0.002	0.576	0.250	0.388	0.433	1.650
8.5	0.000	0.308	0.234	0.274	0.386	1.201
9.5	0.001	0.352	0.223	0.308	0.370	1.254
10.5	0.000	0.188	0.215	0.270	0.356	1.029
11.5	0.000	0.256	0.221	0.252	0.340	1.070
12.5	0.000	0.247	0.183	0.259	0.322	1.012
13.5	0.000	0.183	0.321	0.232	0.414	1.150
14.5	0.000	0.195	0.158	0.277	0.361	0.991
15.5	0.000	0.126	0.164	0.256	0.381	0.927
16.5	0.000	0.160	0.163	0.246	0.306	0.876
17.5	0.000	0.201	0.148	0.224	0.290	0.863
18.5	0.000	0.214	0.169	0.210	0.283	0.877
19.5	0.000	0.261	0.183	0.207	0.273	0.925
20.5	0.000	0.299	0.183	0.224	0.272	0.979
21.5	0.000	0.263	0.129	0.233	0.216	0.843
22.5	0.000	0.310	0.151	0.247	0.266	0.974
23.5	0.001	0.210	0.240	0.338	0.339	1.128
24.5	0.001	0.218	0.132	0.233	0.268	0.852
25.5	0.000	0.233	0.132	0.240	0.269	0.875
26.5	0.000	0.283	0.105	0.293	0.270	0.952
27.5	0.000	0.227	0.127	0.223	0.264	0.840
28.5	0.000	0.202	0.127	0.233	0.291	0.853
29.5	0.000	0.202	0.111	0.213	0.264	0.791
31.0	0.000	0.203	0.129	0.213	0.267	0.812
33.0	0.000	0.216	0.162	0.215	0.236	0.830
35.0	0.000	0.193	0.123	0.205	0.263	0.784
37.0	0.000	0.170	0.158	0.184	0.241	0.754
39.0	0.000	0.178	0.153	0.203	0.272	0.807
41.0	0.000	0.178	0.400	0.211	0.261	1.050
43.0	0.000	0.188	0.150	0.183	0.250	0.771
45.0	0.000	0.186	0.167	0.215	0.292	0.860

1042	Table 9. Dry mass and radiochemical results and chronology for sediment core MEL5 in
1043	Lough Melvin.

Depth, cm	Dry mass, g cm ⁻²	Total Pb- 210, Bq kg ⁻¹	Supported Pb- 210. Bq kg ⁻¹	Unsupported Pb-210, Bq kg ⁻¹	Cs-137, Bq kg ⁻¹	Date	±
0.5	0.0563	309.9	79.36	230.54	102.7	2017	

2.5 0.3526 399.52 81.25 318 4.5 0.7513 248.39 75.5 172	3.27 161.78	2016	2
4.5 0.7513 248.39 75.5 172			. — .
	2.89 349.91	2008	2
6.5 1.2048 176.52 68.09 108	3.43 182.83	3 1994	2
7.5 1.4441 168.22 70 98	3.22 222.3	1980	3
8.5 1.6999 140.66 70.28 70	0.38 266.93	8 1972	3
9.5 1.9669 132.14 68.3 63	8.84 84.25	5 1964	4
11.5 2.5077 109.53 74.04 35	5.49 20.78	1955	5
12.5 2.7842 105.6 76.36 29	9.24 16.84	1932	8
13.5 3.0626 97.28 75.35 21	1.93 9.55	5 1919	11
14.5 3.346 85.27 72.33 12	2.94 4.79	1903	16
15.5 3.6567 88.78 72.37 16	6.41 (1883	26
16.5 3.9952 65.21 72.75 -7	7.54 ()	
24.5 6.8784 79.56 81.54 -	1.98 0)	

1045 Table 10. Dry mass and radiochemical results and chronology for sediment core RAM5 in 1046 Lough Ramor.

Depth,	Dry	Total Pb-	Supported Pb-	Unsupported	Cs-137,	Date	±
cm	mass, g	210, Bq kg ⁻¹	210. Bq kg ⁻¹	Pb-210, Bq kg ⁻¹	Bq kg ⁻¹	Date	—
••••	cm ⁻²	,		,	- 4 - 9		
0.5	0.0517	109.5	47.18	62.32	43.21	2017	2
2.5	0.3361	133.79	41.72	92.07	51.26	2013	2
4.5	0.6815	90.81	42.66	48.15	51.26	2009	2
6.5	1.0492	131.03	48	83.03	60.47	2004	2
8.5	1.4675	90.57	47.09	43.48	69.19	1997	3
10.5	1.9154	94.56	42.43	52.13	68.52	1990	4
12.5	2.378	73.1	40.27	32.83	70.58	1983	5
14.5	2.8713	77.47	42.08	35.39	72.34	1974	6
16.5	3.388	72.66	43.22	29.44	63.4	1962	9
18.5	3.9301	61.07	44.97	16.1	39.29	1950	13
20.5	4.5324	48.24	40.59	7.65	26.68	1941	16
22.5	5.2079	56.41	42.54	13.87	19.07	1926	23
24.5	5.9188	46.66	37.59	9.07	11.68	1893	28
28.5	7.3218	33.21	36.31	-3.1	2.86		

1047

1048 Table 11. Dry mass and radiochemical results and chronology for sediment core SHE5 in 1049 Lough Sheelin.

Depth,	Dry	Total Pb-	Supported Pb-	Unsupported	Cs-137,	Date	±
cm	mass, g	210, Bq kg ⁻¹	210. Bq kg ⁻¹	Pb-210, Bq kg ⁻¹	Bq kg⁻¹		
	cm ⁻²						
0.5	0.0794	0.0794	36.52	123.88	48.22	2018	2
4.5	0.7536	0.7536	37.06	137.66	43.32	2016	2
9.5	1.7029	1.7029	33.18	131.24	48.96	2013	2
13.5	2.527	2.527	33.11	125.79	47.58	2011	2
19.5	3.8062	3.8062	34.52	103.75	63.36	2007	2
24.5	4.8213	4.8213	30.32	111.2	56.68	2004	3
29.5	5.785	5.785	28.73	102.69	67.61	2001	4
34.5	6.798	6.798	28.24	94.89	70.29	1997	4
39.5	7.8394	7.8394	26.3	67.35	80.09	1994	5
44.5	8.9185	8.9185	26.5	83.84	84.71	1991	6
49.5	9.9997	9.9997	24.92	73.22	103.11	1987	8
54.5	11.0482	11.0482	31.99	54.03	107.37	1983	9
57.5	11.6818	11.6818	31.51	45.74	112.72	1981	10
60.5	12.34	12.34	28.3	57.75	116.3	1979	11
62.5	12.7743	12.7743	30.41	58	114.16	1977	11
64.5	13.2078	13.2078	29.63	66.62	116.53	1975	12

Table 12. Dry mass and radiochemical results and chronology for sediment core LN11 in Lough Neagh.

1032

Depth,	Dry	Total Pb-	Supported Pb-	Unsupported	Cs-137,	Date	±
cm	mass, g	210, Bq kg ⁻¹	210. Bq kg ⁻¹	Pb-210, Bq kg ⁻¹	Bq kg⁻¹		
	cm ⁻²						
0.5	0.0655	114.72	34.01	80.71	20.37	2018	2
2.5	0.46	122.69	23.59	99.1	25.27	2011	2
3.5	0.7195	89.25	29.46	59.79	31.6	2006	2
5.5	1.355	73.91	23.19	50.72	37.2	1996	4
7.5	2.0005	65.34	22.65	42.69	64.25	1983	6
9.5	2.5745	41.16	27.26	13.9	59.88	1975	7
11.5	3.1435	46.21	25.15	21.06	22.82	1967	9
13.5	3.742	62.14	26.04	36.1	13.62	1948	16
14.5	4.0655	47.39	29.79	17.6	10.05	1932	26
15.5	4.399	38.75	26.81	11.94	9.63	1916	30
16.5	4.7565	28.35	27.91	0.44	4.03		
17.5	5.1255	40.53	23.3	17.23	10.2		
18.5	5.4845	24.72	23.88	0.84	6.03		
19.5	5.863	27.08	22.53	4.55	2.99		
21.5	6.6345	26.84	25.82	1.02	1.74		

Table 13. Dry mass and radiochemical results and chronology for sediment core LN15 in Lough Neagh

Lough N	_ough Neagh.							
Depth,	Dry	Total Pb-	Supported Pb-	Unsupported	Cs-137,	Date	±	
cm	mass, g	210, Bq kg ⁻¹	210. Bq kg ⁻¹	Pb-210, Bq kg ⁻¹	Bq kg⁻¹			
	cm ⁻²							
0.5	0.0489	149.61	34.73	114.88	48.75	2019	2	
4.5	0.6169	163.7	38.71	124.99	64.62	2013	2	
8.5	1.3057	141.3	34.9	106.4	73.31	2005	2	
12.5	2.1187	114.01	28.76	85.25	114.49	1995	3	
16.5	3.0554	68.73	33.76	34.97	137.83	1986	4	
20.5	4.0439	59.23	30.96	28.27	71.63	1979	5	
23.5	4.747	60.14	28.9	31.24	98.65	1974	6	
27.5	5.6671	46.85	28.52	18.33	36.26	1966	7	
31	6.519	67.93	26.94	40.99	15.65	1956	10	
35	7.5116	42.14	25.96	16.18	9.14	1938	16	
39	8.5201	29.38	26.79	2.59	2.12	1928	19	
43	9.5445	32.88	25.52	7.36	2.63	1924	21	
47	10.5815	47.09	32.86	14.23	0	1903	31	
51	11.6305	34.72	28.64	6.08	0			
55	12.6787	31.03	34.01	-2.98	0			

Table 14. Dry mass and radiochemical results and chronology for sediment core LN17 in Lough Neagh.

1058

Lought	Lough Neagh.								
Depth,	Dry	Total Pb-	Supported Pb-	Unsupported	Cs-137,	Date	±		
cm	mass, g cm ⁻²	210, Bq kg ⁻¹	210. Bq kg ⁻¹	Pb-210, Bq kg⁻¹	Bq kg⁻¹				
0.5	0.0489	149.61	34.73	114.88	48.75	2019	2		
4.5	0.6169	163.7	38.71	124.99	64.62	2013	2		
8.5	1.3057	141.3	34.9	106.4	73.31	2005	2		
12.5	2.1187	114.01	28.76	85.25	114.49	1995	3		
16.5	3.0554	68.73	33.76	34.97	137.83	1986	4		
20.5	4.0439	59.23	30.96	28.27	71.63	1979	5		

23.5	4.747	60.14	28.9	31.24	98.65	1974	6
27.5	5.6671	46.85	28.52	18.33	36.26	1966	7
31	6.519	67.93	26.94	40.99	15.65	1956	10
35	7.5116	42.14	25.96	16.18	9.14	1938	16
39	8.5201	29.38	26.79	2.59	2.12	1928	19
43	9.5445	32.88	25.52	7.36	2.63	1924	21
47	10.5815	47.09	32.86	14.23	0	1903	31
51	11.6305	34.72	28.64	6.08	0		
55	12.6787	31.03	34.01	-2.98	0		

Table 15. Dry mass and radiochemical results and chronology for sediment core LN18 in Lough Neagh.

Loughin	eagn.						
Depth,	Dry	Total Pb-	Supported Pb-	Unsupported	Cs-137,	Date	±
cm	mass, g	210, Bq kg ⁻¹	210. Bq kg ⁻¹	Pb-210, Bq kg ⁻¹	Bq kg⁻¹		
	cm ⁻²						
0.5	0.0555	182.48	41.7	140.78	60.71	2017	2
4.5	0.711	137.19	37.31	99.88	63.81	2011	2
9.5	1.7235	122.05	41.77	80.28	79.95	2002	2
12.5	2.3785	134.22	34.53	99.69	119.97	1993	2
14.5	2.8285	100.23	36.16	64.07	165.69	1987	3
17.5	3.5285	103.81	32.72	71.09	271.52	1977	4
19.5	4.0605	68.05	31.33	36.72	95.01	1968	5
22.5	4.86	65.22	27.08	38.14	83.07	1956	7
24.5	5.393	52.06	29.56	22.5	39.04	1947	9
26.5	5.992	32.55	26.37	6.18	18.86	1940	10
28.5	6.6345	51.72	30.8	20.92	11.44	1935	12
31	7.386	36.57	33.26	3.31	18.06	1929	14
33	7.9648	41.19	33.37	7.82	10.16	1923	15
37	9.1268	51.8	33.52	18.28	17.93	1890	28
41	10.2808	29.77	34.44	-4.67	1.56		

Table 16. Dry mass and radiochemical results and chronology for sediment core LN19 in Lough Neagh.

1064	
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Lough Neagh.							
Depth,	Dry	Total Pb-	Supported Pb-	Unsupported	Cs-137,	Date	±
cm	mass, g	210, Bq kg ⁻¹	210. Bq kg ⁻¹	Pb-210, Bq kg ⁻¹	Bq kg⁻¹		
	cm ⁻²						
0.5	0.06	138.13	27.38	110.75	67.73	2017	2
2.5	0.3535	151.28	27.69	123.59	85.08	2008	2
4.5	0.707	101.97	30.99	70.98	108.11	1995	3
5.5	0.893	101.35	28.36	72.99	129.96	1987	3
6.5	1.0805	100.92	32.11	68.81	157.75	1978	4
7.5	1.263	74.3	33.2	41.1	150.2	1969	5
8.5	1.436	66.68	27.97	38.71	103.73	1960	6
9.5	1.617	64.97	28.7	36.27	48.49	1948	8
10.5	1.814	50.65	27.6	23.05	38.06	1934	12
11.5	2.0195	36.38	27.16	9.22	20.17	1922	17
12.5	2.2315	49.26	29.76	19.5	22.37	1904	26
13.5	2.4475	34.9	29.67	5.23	11.39	1874	33
15.5	2.9185	31.77	32.27	-0.5	13.81		
16.5	3.1705	34.84	29.93	4.91	4.82		

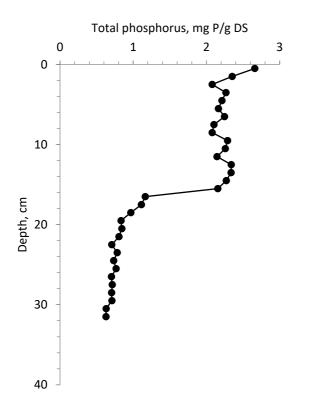




Fig. 1. The variation of total phosphorus with depth in sediment core SHE5 in LoughSheelin, retrieved on 16 May 2018 from 8.7 m water depth.

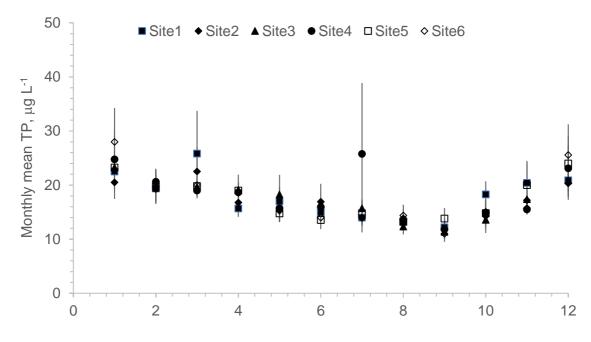


Fig. 2. The seasonal variation of monthly mean (±SE) total phosphorus concentration at six
sites in Lough Melvin over the 2007 to 2019 period.

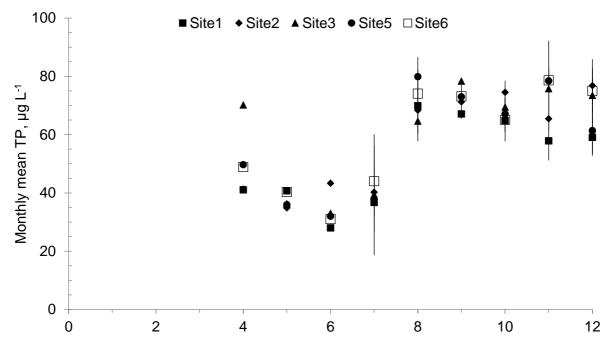




Fig. 3. The seasonal variation of monthly mean (±SE) total phosphorus concentration at five
sites in Lough Ramor over the 2007 to 2016 period.

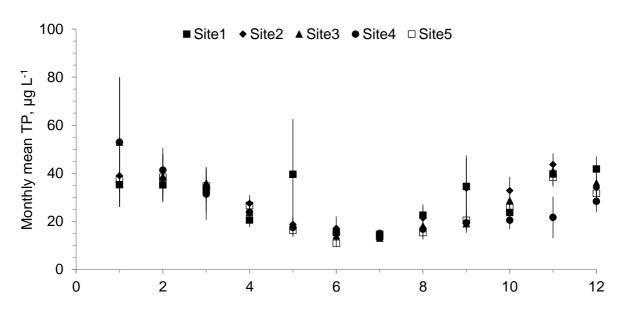


Fig. 4. The seasonal variation of monthly mean (±SE) total phosphorus concentration at five
sites in Lough Sheelin over the 2007 to 2016 period.